

Spatial variability of soil properties and soil erodibility

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Spatial variability of soil properties and soil erodibility in the Alqueva dam watershed, Portugal

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Abstract

The aim of this work is to investigate how the spatial variability of soil properties and soil erodibility (K factor) were affected by the changes in land use allowed by irrigation with water from a reservoir in a semiarid area. To this, three areas representative of different land uses (agroforestry grassland, Lucerne crop and olive orchard) were studied within a 900 ha farm. The interrelationships between variables were analyzed by multivariate techniques and extrapolated using geostatistics. The results confirmed differences between land uses for all properties analyzed, which was explained mainly by the existence of diverse management practices (tillage, fertilization and irrigation), vegetation cover and local soil characteristics. Soil organic matter, clay and nitrogen content decreased significantly, while K factor increased with intensive cultivation. The HJ-biplot methodology was used to represent the variation of soil erodibility properties grouped in land uses. Native grassland was the least correlated with the other land uses. K factor demonstrated high correlation mainly with very fine sand and silt. The maps produced with geostatistics were crucial to understand the current spatial variability in the Alqueva region. Facing the intensification of land-use conversion, a sustainable management is needed to introduce protective measures to control soil erosion.

1 Introduction

Soil erosion is a significant economic and environmental problem worldwide as a driving force affecting landscape (Zhao et al., 2013). It is a very dynamic and complex process, characterized by the decline of soil quality and productivity, as it causes the loss of topsoil and increases runoff (Lal, 2001; Yang et al., 2003). Furthermore, soil erosion often causes negative downstream impacts, such as the sedimentation in rivers and reservoirs decreasing their storage volume as well as lifespan (Pandey et al., 2007; Haregeweyn et al., 2013).

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One of the main cause of soil loss intensification around the world is associated with land-use change (Leh et al., 2013). The relationship between different land use and soil susceptibility to erosion has attracted the interest of a variety of researchers (Yang et al., 2003; Cerdà and Doerr, 2007; Blavet et al., 2009; Biro et al., 2013; Wang and Shao, 2013), which have shown the impact of changes on vegetation cover and agricultural practices on soil properties and therefore in overland flow. Generally, cultivated lands experience the highest erosion yield (Cerdà et al., 2009; Mandal and Sharda, 2013). In the Mediterranean regions, in combination with these anthropogenic factors, the climate change has amplified the concerning about soil erosion since it is expected the increase of dry periods followed by heavy storms with concentrated rainfall (Nunes et al., 2009).

Some models have been developed to predict soil loss and sediment delivery. The Revised Universal Soil Loss Equation (RUSLE) is the most used empirical equation for modeling annual soil loss from agricultural watersheds (Renard et al., 1997). The susceptibility of soil erosion and land degradation depends largely on various inherent soil properties, namely chemical, physical, biological and mineralogical properties (Cambardella et al., 1994; Pérez-Rodríguez et al., 2007). However, according to the RUSLE model only some of the soil's properties define soil erodibility (K factor), such as particle-size composition, the content of organic matter, soil structure and permeability. Therefore, the K factor is the most used and is an important index to measure soil susceptibility to erosion (Panagopoulos and Antunes, 2008).

Spatial variability in soils occurs naturally as a result of complex interactions between geology, topography and climate. Moreover the spatial variability of soil properties, which influence soil susceptibility to erosion, is highly related with anthropogenic factors particularly in cultivated lands (Paz-González et al., 2000; Wang and Shao, 2013). Then, information on the spatial variability and the interactions between soil properties is essential for understanding the ecosystem processes and planning sustainable soil management alternatives for specific land-uses (Pérez-Rodríguez et al., 2007; Ziadat and Tamimeh, 2013).

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Classical statistics and geostatistics methods have been widely applied on studies about spatial distribution of soil properties (Pérez-Rodríguez et al., 2007; Tesfahunegn et al., 2011). Geostatistical techniques based on predictions and simulations have been used to describe areas where predicted information is established by a limited number of samples (Goovaerts, 1997). Geostatistics provides tools for analyzing spatial variability structure and distribution of soil properties and evaluating their dependence (Panagopoulos et al., 2014).

The Biplot methodology provides an added value for analyzing spatial variability of soil properties. This multivariate statistical technique allows the graphical representation of a large data matrix (Gabriel, 1971), whereby it is possible to interpret the relations between individuals (samples) and between variables, as well as between both. Biplot can also indicate clustering of units with close characteristics, showing inter-unit distances as well as displaying variances and correlations of the variables (Gallego-Álvarez et al., 2013). The HJ-Biplot permits not only the analysis of the behavior by sample but also the determination of which variable is responsible for such behavior (García-Talegón et al., 1999), allowing a visual appraisal to establish relations between soil properties and land uses.

The construction of the Alqueva dam in a semiarid area of South Portugal created one of the largest artificial lakes in Europe. Taking advantage of water availability from the reservoir, this Mediterranean region has been subjected to land-use conversion from the native Montado grassland to intensive agricultural uses. Land-use conversion from the native ecosystem to agriculture may alter physical, chemical and biological soil properties which consequently may increase soil erosion and siltation in the reservoir. Soil erosion in the area has to be carefully evaluated in order to take sustainable soil management measures. Therefore, the aim of this study was to evaluate the effects of cultivation practices on some chemical and physical soil properties and on soil erodibility (K factor on RUSLE), and to characterize their spatial variability using geostatistics and HJ-Biplot methodology.

2 Material and methods

2.1 Study area

Localized in the semiarid Alentejo region of Portugal, at the Guadiana River, the Alqueva reservoir (8°30' W, 38°30' N) covers an area of 250 km², and the capacity of the reservoir is 4.15 km³. The main arguments for the implementation of what is considered the largest artificial lake in Europe were based on the need to combat the growing effects of desertification and to prevent the annual and monthly fluctuations in precipitation supply. One of the main goal of the Alqueva Multipurpose Project was the implementation of 120 000 ha of new irrigated land in the Alentejo. The Alentejo region, covering an area of 27 000 km² is considered one of the most depressed regions of the European Union and characterized by a Mediterranean climate with very hot and dry summers and mild winters. The average temperature ranges from 24 to 28 °C in hot months (July/August) and from 8 to 11 °C in cold months (December/January). The average annual precipitation at the nearest meteorological station, for the last 30 years, is 517.2 mm. The region is affected by intense dry periods followed by heavy, erosive rains concentrated in the autumn season.

The study experimental site (farm “Herdade dos Gregos”), located in the surrounding area of the reservoir (Fig. 1), is a private property with 900 ha. The landscape is characterized by its hilly topography with significant altitude variations (mainly between 100 and 250 m). The bedrock of the study area is rocky and according to World Reference Base for Soil Resources (FAO, 2006), the two types of soil in this area are: haplic luvisols (LVha) and Lithic leptosols (LPli). This farm was selected to include a diversity of land uses, including native Montado grassland and more intensive land-uses, with irrigation, namely Olive tree orchard and Lucerne cultivation. Direct pumping from Alqueva reservoir is done in this private property since it is near the reservoir.

The typical landscape in the Alentejo region is the Montado native grassland, an agrosilvopastoral system characterized by savannah-like, low density woodlands with evergreen holm oaks (*Quercus ilex*). For that reason, an area of the Montado grassland

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(20.7 ha), used as a permanent pasture for the cattle, was selected for this study. This small area is located in the high altitudes of the “Herdade dos Gregos” (from 200 to 240 m) with a slope that varies from 1.4 to 20.9%. Tillage (at about 15 cm depths) was done only once every 10 years to decrease shrub competition (the most recent one was four years before the study implementation), and the soil is not subjected to any fertilizer. Four years before the study implementation, there was a fire on this agrosilvopastoral area of the farm.

Taking advantage of the water availability, another land use (with 33.5 ha) is an irrigation area (Pivot Sprinkler Irrigation System) on which Lucerne (*Medicago sativa*) is sown four times a year. Lucerne, once dried, is nutritional for cattle, and it incorporates nitrogen in the soil. In this area, conventional tillage is used, involving multiple aspects: plough (about 20 cm depth) in fall, following cultivator (about 15 cm depths) and disc harrow (about 10 cm depths) subsequent to soil tillage. Inorganic fertilizers were applied to the cultivated field at a rate of 100 kg N : P : K ha⁻¹. This land use is placed in the midland (194–220 m), and the slope varies from 0 to 9%.

Other irrigated land use consists of an Olive tree plantation (57.5 ha), which is done in strips. This cultivation has a drip irrigation system, is fertilized once every two years and is ploughed once a year to decrease weed competition. The Olive orchard is located in the low elevations of the farm (150–186 m), and it is on the side of the reservoir (Fig. 1). The slope varies from 0 to 14.2%.

2.2 Soil sampling and laboratory analysis

Since the objective was to study the relation between soil properties and *K* factor from RUSLE, the soil samples were collected from 0 to 20 cm depth, according to Renard et al. (1997). In order to predict variations in short distances, 25, 27 and 52 soil samples were randomly collected respectively in Montado, Lucerne and the Olive orchard (see Fig. 1). Samples were air-dried and then dried for about 6 h at 40 °C on a ventilated oven, and they were passed through a 2 mm sieve to remove rocks or limestone concretions. The particle-size distribution was determined by the Bouyoucos

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hydrometer method (Bouyoucos, 1936). Soil organic matter content was determined using the Walkley and Black (1934) method, a wet oxidation procedure. The soil's total nitrogen content was determined according to Kjeldhal digestion, distillation and the titration method (Bremmer and Mulvaney, 1982). Soil pH and electrical conductivity were measured with glass electrode in a 1 : 2.5 soil/water suspension (Watson and Brown, 2011).

2.3 Soil erodibility factor

Soil erodibility factor (K) ($\text{Mg ha h ha}^{-1} \text{ MJ}^{-1} \text{ mm}^{-1}$) was estimated using soil property values, such as particle-size composition, content of organic matter, soil structure and permeability, in the 104 samples points described above. This factor represents the soil-loss rate per erosion index unit for a specified soil as measured on a standard plot (Renard et al., 1997). An algebraic approximation of the nomograph was used to estimate K factor (Renard et al., 1997):

$$K = [2.1 \times 10^{-4}(12 - \text{OM}) \times M^{1.14} + 3.25(s - 2) + 2.5(p - 3)]/759 \quad (1)$$

where OM is the percentage of organic matter, s is soil structure class, p is permeability class, and M is the product of the percentage of modified silt (silt particles and very fine sand) or the 0.002–0.1 mm size fraction and the sum of the percentage of silt and percentage of sand. K is expressed with SI units of $\text{Mg ha h ha}^{-1} \text{ MJ}^{-1} \text{ mm}^{-1}$. To estimate the permeability the field-saturated hydraulic conductivity was measured in the field using a double-ring infiltrometer (6 site-measurements per land-use, each one with 5 repetitions). Permeability class and soil structure class were defined in accordance with Renard et al. (1997).

2.4 Statistical and Geostatistical Analysis

Data were subjected to classical analysis using SPSS 17.0 software to obtain descriptive statistics, namely the mean, minimum and maximum, SD, coefficient of variation (CV) and skewness of each parameter.

Geostatistical analysis to examine spatial distribution of soil properties were performed in ArcGIS 10 (Geostatistical Analyst Tool). Prior to geostatistics to obtain prediction maps, data transformation to normal distribution was necessary for some soil properties, using geostatistical analyst tools (log or box-cox method). Skewness is the most common statistic parameter to identify a normal distribution that is confirmed with skewness values varying from -1 to $+1$. Trend analysis was performed to examine the presence of any global directional trend in our data, an overriding process that affects all measurements in a deterministic way (nonrandom). So, when necessary, the trend removal was done using Geostatistical Analyst tools to more accurately model the variation (Panagopoulos et al., 2006).

The semivariogram (SV) is the graphical representation that determines how samples are related to each other in space and it is based on:

$$\gamma(h) = 1/2N(h) \times \sum [Z_i - Z_{(i+h)}]^2 \quad (2)$$

where $\gamma(h)$ is the variance (the most related samples have lower values of variance), $N(h)$ is the number of samples that can be grouped using vector h , Z_i represents the value of the sample, and Z_{i+h} is the value of another sample located at a distance $\|h\|$ from the initial sample Z_i (Chiles and Delfiner, 1999).

The spatial structure and dependence of each variable has been described by semivariogram parameters. Nugget is the variance at distance zero and reflects the sampling error. Sill is the semivariance value at which the semivariogram reaches the upper bound and flattens out after its initial increase; it is the variance in which the samples are no longer spatially related at the study area.

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The semivariogram was calculated for all the measured properties. Semivariogram models were selected by employing the cross-validation technique, which permits the evaluation of the prediction accuracy. Cross-validation compares statistical values (as the mean error [ME] or root-mean-square standardized error [RMSSE]) estimated from the models and real values.

2.5 HJ-Biplot

HJ-Biplot represents a matrix, without assumptions related to its probabilistic distribution, permitting a graphic representation of the geometric data structure, representing the dataset (samples and variables) variability. The prefix “bi” is due to a simultaneous representation of the matrix rows and columns, searching for the maximum representation quality possible, at the same scale (Martín-Rodríguez et al., 2002; González-Cabrera et al., 2006; Gallego-Álvarez et al., 2013).

A data matrix \mathbf{X} suffers a factorization to reduce its dimensionality through single value decomposition, the algebraic base of biplot representation (Eq. 3) (Gabriel, 1971).

$$\mathbf{X}_{(n \times p)} = \mathbf{U}_{(n \times r)} \mathbf{\Lambda}_{(r \times r)} \mathbf{V}'_{(r \times p)} \quad (3)$$

where $\mathbf{\Lambda}_{(r \times r)}$ is a diagonal $(\lambda_1, \lambda_2, \dots, \lambda_r)$ corresponding to the r eigenvalues of $\mathbf{X}\mathbf{X}'$ or $\mathbf{X}'\mathbf{X}$, $\mathbf{U}_{(n \times r)}$ is an orthogonal matrix whose columns are the eigenvectors of $\mathbf{X}\mathbf{X}'$, and $\mathbf{V}'_{(r \times p)}$ is an orthogonal matrix whose columns are the eigenvectors of $\mathbf{X}'\mathbf{X}$.

With the *MultiBiplot software*, developed by the University of Salamanca (Vicente Villardón, 2014), an HJ-Biplot was used to determine the relation between soil properties, between land uses, and the correlations between both (soil properties and land uses), thereby defining patterns and clustering the samples in groups.

On this type of graphic representation, the points represent individuals (samples), and the vectors represent variables (in this case, chemical and physical soil properties). The interpretation of the variables is based on the cosine of the angles between

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soil type. The Lithic leptosols (LPLi) soils are characterized by a thin layer (about 10 cm), in that case upon a schist rock, justifying the higher clay content at the Montado grassland. The Haplic luvisols (LVha) soils in the Lucerne cultivation and the Olive orchard are characterized by a loam or sandy loam layer (first 20 cm) with good drainage over clay-enriched subsoil (upon a basic crystalline rock), explaining the lower values of clay and fine sand, especially in the Olive orchard. Despite the same soil type, soil texture is different between Lucerne and Olive orchard that can be justified by land-use. The Lucerne is a more intensive cultivation (intensive irrigation, tillage and continuous cultivation, fertilizers and lime application), conditions that promote changes in the soil weathering and moisture, and consequently on soil texture (Yimer et al., 2008). On the other hand the soil between olive trees is kept without vegetation for most of the year and it can explain the clay drainage to a sub-layer.

Montado shows the highest content of OM (5.22 %), whereas Lucerne and Olive fields show the lowest values (with 2.08 and 2.10 %, respectively). Other studies suggest that OM is higher in no-tillage soils compared to minimum tillage that increases aeration (Celik, 2005). Tillage mixes the subsoil with topsoil; after soil erosion, the nutrients are easily leached and the surface becomes poor in nutrients (Al-Kaisi and Licht, 2005). As for OM, the highest values of N nutrient occur in the Montado (0.19 %) and the lowest values in Lucerne (0.11 %) and the Olive orchard (0.10 %), which is related to the tillage practice that is frequently employed in these last two land uses, while in the Montado grassland the cattle enriches the soil.

Soil EC values (Table 1) were similar when comparing the Montado grassland (0.100 dS cm^{-1}) and the Lucerne field (0.107 dS cm^{-1}); they were slightly higher in the Olive orchard (0.182 dS cm^{-1}) but not enough to raise salinity problems. Usually, the addition of fertilizers (that happens on Lucerne and the Olive orchard) can cause high EC due to the percentage of the salts, which are leached by water irrigation (higher in the Lucerne field).

The soil pH was significantly higher in the Lucerne cultivated land (7.1) compared to the Montado grassland (5.9) or in the Olive tree orchard (5.5) (Table 1). The soil

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pH in the Lucerne was greater due to lime application to increment the soil pH in that area. Lucerne's optimum pH for production is between 6.5 and 7.2, and lime application has been found to produce a significant improvement in nodulation of Lucerne (both number and dry weight of nodules per plant) (Grewal and Williams, 2001). According to Chatterjee and Lal (2009), minimum tillage soils had higher soil pH values than plow tillage soils; this was verified when comparing the Montado grassland, with no tillage, and the olive orchard, with tillage between tree lines.

Saturated Hydraulic Conductivity (HC) values were greater in the Lucerne area (5.95 cm h^{-1}), slightly lower in the Montado grassland (4.56 cm h^{-1}) and lowest in the Olive orchard (2.60 cm h^{-1}). The lower permeability in the Olive orchard can be explained by the clay-enriched subsoil or soil crust problems, and may explain the higher values of EC, i.e. the greater concentration of salts. Also it can be explained by the frequency of tillage in the different land uses because aggregate stability and water infiltration rate are higher in soils subjected to limited tillage systems (Alvarez and Steinbach, 2009).

As a result, K factor was different for the typical land use, Montado grassland, compared to the Lucerne cultivation and the Olive orchard. The values increased with the intensification of the cultivation field, with the lowest values for Montado grassland ($0.021 \text{ Mg ha h ha}^{-1} \text{ MJ}^{-1} \text{ mm}^{-1}$) and the highest for the Lucerne cultivation ($0.039 \text{ Mg ha h ha}^{-1} \text{ MJ}^{-1} \text{ mm}^{-1}$) and the Olive orchard ($0.038 \text{ Mg ha h ha}^{-1} \text{ MJ}^{-1} \text{ mm}^{-1}$). Other studies had similar results, showing that the removal of permanent vegetation, the loss of OM and the reduction of aggregation, caused by intensive cultivation, contribute to decrease K factor (Celik, 2005).

3.2 Spatial dependence of soil properties

Model selection for each soil property was based on the nugget, sill, mean error (ME) and the root-mean-square standardized error (RMSSE) presented in the second part of Table 1 (Geostatistics).

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Looking for natural vs. anthropogenic impact on the K factor, for each land-use, it's evident that in the Montado the spatial variability is mainly associated with natural (intrinsic) factors (as texture), being soil properties and erodibility distribution more homogenous. In the Lucerne and Olive orchard the spatial variability is more dependent from not homogenous anthropogenic causes such as fertilization and irrigation rates and tillage/plough processes.

3.4 HJ-Biplot

The HJ-Biplot representation matrix of soil properties is shown in Fig. 4. It was observed that the dominant axis (axis 1) takes 35.83% of the total inertia (information) of the system. With both dimensions, an accumulative inertia of 61.04% was achieved. Regarding this graphic representation, it was observed that samples were grouped according to the land use. The Montado samples were close to OM, N and Clay vectors, showing their preponderance to be a characterization of these variables. The Lucerne samples were important to describe the pH and Silt content. On the other hand the Olive samples were more disperse but related to EC, Permeability class, Sand, VFS and K .

The variables demonstrating a more positive correlation between them were OM and N, as previously noticed. Clay and Silt were also positively correlated, but negatively correlated with sand as expected, because soils with more sand have less clay and/or silt.

Through the matrix representation it was detected that soils with more sand have higher EC (Olive orchard), although EC normally increases with the percentage of clay. This may be explained by the addition of fertilizers, as previously discussed, that can contribute to an EC increase. These results for EC show low variability between land uses, revealing a low cation exchange capacity (CEC) of these soils. This is frequently caused by intensive soil mobilization (Paz-González et al., 2000).

Permeability class increases as the HC_{sat} decreases, as defined by Renard et al. (1997). So, contrary to what was expected, for this study the soils with more

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sand (occurring in the Olive orchard) have less hydraulic conductivity (high permeability class). It can be explained by a clay-enriched sub-layer under the sandy loam layer or/and by the soil compaction/degradation processes. The soil compaction and degradation can be related to repeated plow operations to reduce shrubs between olive rows and irrigation (Pagliai et al., 2004). This permeability decrease in the Olive orchard was correlated with the increase of K factor.

Nevertheless, the properties more positively correlated with K factor were the very fine sand (VFS) and silt; this is due to the susceptibility of these particles to erosion since they can be easily detached and transported by water (Morgan, 2005). The OM and N content were negatively correlated with K and permeability. The higher OM reduces the susceptibility of the soil to detachment and increases infiltration (Bronick and Lal, 2005). The nitrogen (N) content is not used to estimate K ; however, especially for soils without fertilization, the existent N is mostly associated to OM. Nevertheless, nutrients decrease in soils that are more erodible, according to the literature (Tesfahunegn et al., 2011). The clay content also shows a negative correlation with K factor, as expected (Renard et al., 1997).

Figure 5 shows the hierarchical clusters representation. Using HJ-Biplot methodology and the aggregation tool *ward*, 3 clusters were obtained. The samples were grouped by land uses (that were already detected by the matrix representation, see Fig. 4). *Cluster 1* is represented by a majority of samples from Lucerne, *Cluster 2* by samples from Montado and *Cluster 3* by samples from the Olive orchard. This was explained by the effect of different management practices, vegetation cover and local soil characteristics, as discussed. Some samples in each land use had different values (higher or lower than the majority) and were grouped in a different *cluster*. Identifying the location of the sample, the cause of displacement can be studied and can help to improve land management practices.

4 Conclusions

This study demonstrated that the variability of soil properties and K factor is associated to land use, cultural practices (tillage type, fertilizer rates, conservation measures, etc.) and local conditions (complex topographic landscape, soil type, etc.). The K factor showed high correlation especially with organic matter, nitrogen, silt and very fine sand. Soils with intensively cultivated land use, and consequently with more tillage and irrigation, had lower organic matter and lower nitrogen content. This translates into a lower cation exchange capacity producing lower aggregate stability and, consequently, an increase of the K factor.

Therefore, in the surrounding area of the Alqueva reservoir, the ongoing change in land use and soil management practices can have a significant effect for chemical and physical soil properties. As a result, this affects the soil erodibility index, intensifying the risk of erosion. The increase of soil loss in the watershed might have a significant impact on a reservoir's ability to storage water, reducing its lifespan.

Knowledge of soil spatial variability is fundamental for environment management and can help in the sustainable use of the resource soil. The prediction maps produced with geostatistics are an important monitoring tool, showing the exact position in the field of the specific soil properties. The simultaneous utilization of the HJ-Biplot methodology was demonstrated to be useful in gaining a better understanding of the spatial variability. This allowed not only a determination of the behavior by sample but also a conclusion as to which variable is responsible for such behavior. Facing the intensification of cultivation in the surrounding area of the reservoir, site-specific soil management and careful land use planning are needed to take into account the spatial variability of soil properties. This area requires special attention on tillage, application rate and type of fertilizers, irrigation scheduling, conservation practices and other efforts.

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**Table 1.** Descriptive statistics of soil properties and parameters of the fitted variogram models and the cross validation results.

	Classic Statistics					Geostatistics					
	Mean	CV (%)	Min	Max	Skewness	Variogram	Nugget	Sill	Nugget/Sill	ME	RMSSE
Montado grassland (<i>n</i> = 25)						Montado grassland (<i>n</i> = 25)					
Clay (%)	17.29	37.7	5.68	29.62	0.07	Exponential	0	38.30	0.00	0.0055	1.01
Silt (%)	29.55	17.2	12.99	39.72	-0.99	Exponential	0	36.00	0.00	0.0238	1.04
Sand (%)	53.16	13.5	39.68	70.34	0.33	Pentaspherical	0	57.60	0.00	0.0223	0.99
VFS (%)	11.13	25.6	4.49	19.04	0.16	Stable	0	12.00	0.00	-0.0188	0.99
OM (%)	5.22	32.1	2.25	10.35	1.19	Exponential*	0.031	0.07	0.44	-0.0003	1.04
N (%)	0.19	43.2	0.07	0.42	1.13	Exponential*	0.056	0.17	0.32	0.0001	1.04
EC (dSm ⁻¹)	0.100	38.1	55.5	217.5	1.28	Exponential*	0.012	0.13	0.09	0.5640	0.95
pH	5.90	4.2	5.38	6.30	0.01	Exponential	0	0.06	0.00	0.0022	0.99
HC _{sat} (cmh ⁻¹)	4.56	42.9	1.20	7.20	-0.57	-	-	-	-	-	-
<i>K</i> (thahha ⁻¹ MJ ⁻¹ mm ⁻¹)	0.021	31.4	0.006	0.039	0.43	Stable	0	0.001	0.00	0.0001	1.00
Lucerne cultivation (<i>n</i> = 27)						Lucerne cultivation (<i>n</i> = 27)					
Clay (%)	13.29	28.8	5.65	22.28	0.32	Stable	0	15.30	0.00	0.0017	1.02
Silt (%)	33.79	26.6	8.35	47.29	-1.48	Stable	0	44.20	0.00	0.0073	0.97
Sand (%)	52.93	17.7	39.32	79.99	1.00	Exponential	0	92.00	0.00	0.0297	0.98
VFS (%)	15.28	37.0	2.59	25.17	-0.39	Exponential	15.60	25.0	0.62	0.0347	1.04
OM (%)	2.08	52.8	0.45	5.44	1.21	Exponential*	15.90	119	0.13	0.0036	0.94
N (%)	0.11	70.2	0.02	0.35	1.43	Circular*	0.10	0.52	0.20	0.0017	1.01
EC (dSm ⁻¹)	0.107	45.9	40.5	205.0	0.64	Exponential	1.15	1.79	0.64	0.2240	0.96
pH	7.14	4.3	6.53	7.85	0.02	Exponential	0.04	0.07	0.57	0.0052	1.07
HC _{sat} (cmh ⁻¹)	5.95	26.7	0.65	1.30	-0.29	-	-	-	-	-	-
<i>K</i> (thahha ⁻¹ MJ ⁻¹ mm ⁻¹)	0.039	21.9	0.013	0.052	-0.88	Stable	0	0.01	0.00	0.0001	1.03
Olive tree orchard (<i>n</i> = 52)						Olive tree orchard (<i>n</i> = 52)					
Clay (%)	9.83	28.8	5.40	16.66	0.52	Stable	0	8.04	0.00	0.0001	0.99
Silt (%)	24.37	46.8	3.82	43.36	-0.41	Pentaspherical	50.00	89.80	0.55	0.0001	0.90
Sand (%)	65.81	18.2	40.6	89.66	0.21	Exponential	0	16.10	0.00	0.0002	0.91
VFS (%)	18.14	32.5	4.49	19.04	0.16	Exponential	0.01	33.70	0.00	0.0037	1.05
OM (%)	2.10	52.8	0.62	8.35	3.54	Exponential*	0.07	0.16	0.44	-0.0006	1.02
N (%)	0.10	45.3	0.04	0.29	2.02	Exponential*	0.02	0.15	0.12	0.0028	1.10
EC (dScm ⁻¹)	0.182	61.3	53.50	583.50	1.80	Exponential	0	1.4	0.00	0.6820	1.02
pH	5.48	7.6	4.30	6.21	-0.43	Exponential	0	0.21	0.00	-0.0002	0.95
HC _{sat} (cmh ⁻¹)	2.60	64.9	0.00	0.67	-0.45	-	-	-	-	-	-
<i>K</i> (thahha ⁻¹ MJ ⁻¹ mm ⁻¹)	0.038	33.6	0.012	0.061	-0.36	Exponential	0.00	0.001	0.51	-0.0001	0.92

* Transformation for normal distribution.

CV – Coefficient variation; Min – minimum; Max – maximum; VFS – Very fine sand; N – Nitrogen; OM – Organic matter; EC – Electrical conductivity; HC_{sat} – Saturated hydraulic conductivity; *K* – Soil erodibility; ME – Mean error; RMSSE – Root-mean-square standardized error

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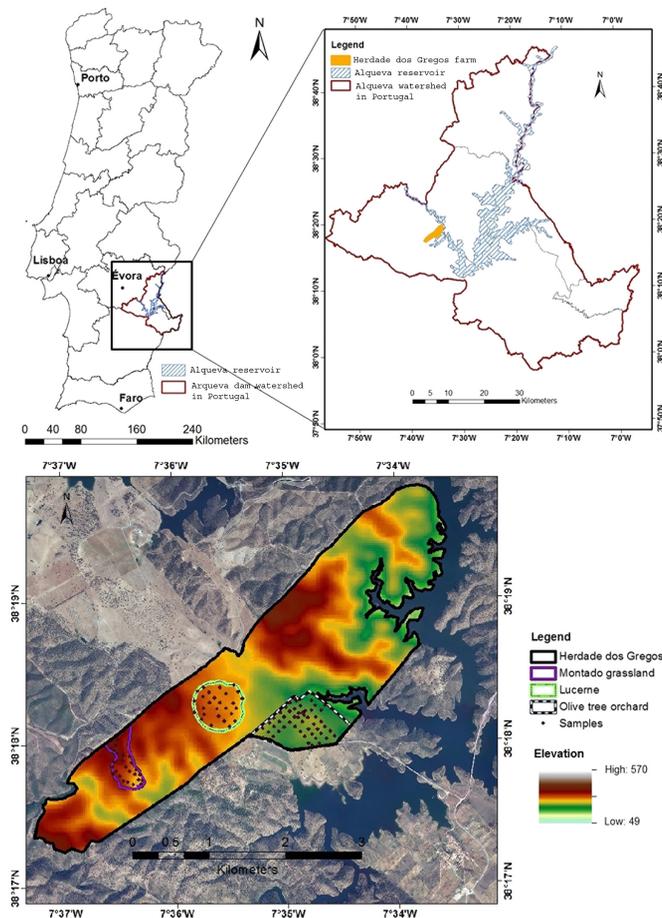


Figure 1. Location of the study area at the Alqueva dam watershed in Portugal.

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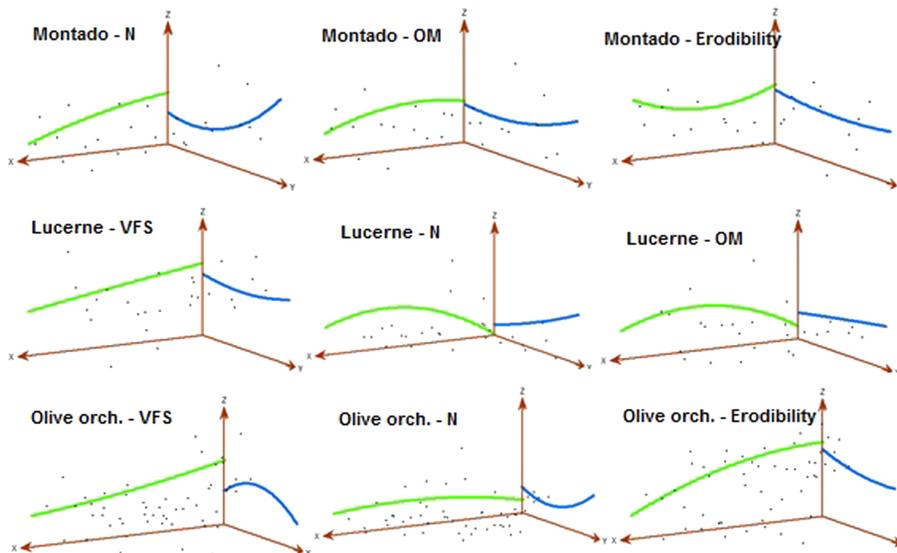


Figure 2. Three-dimensional perspective of the trends in the input datasets.

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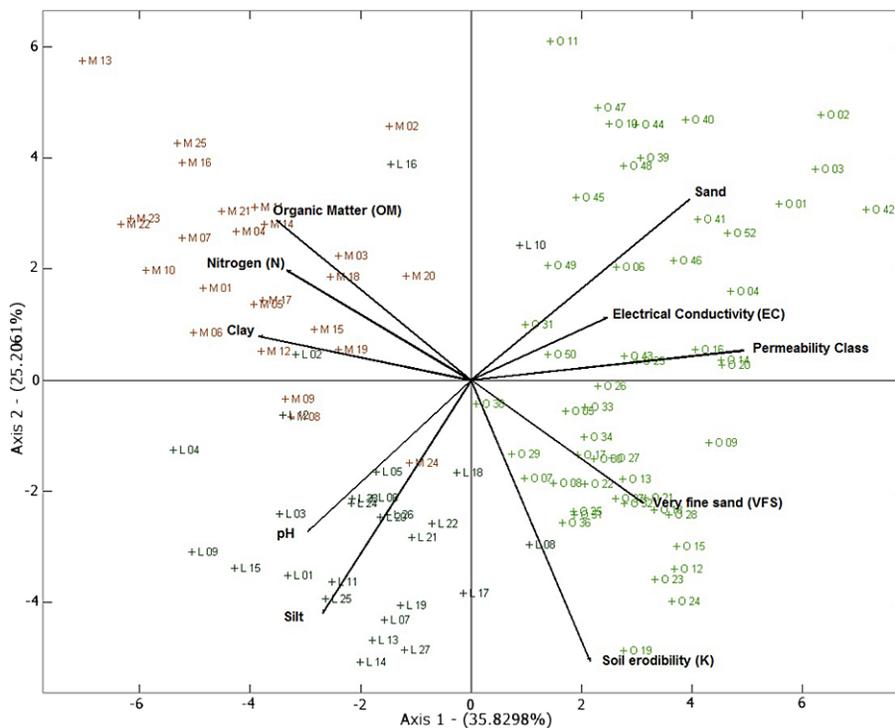


Figure 4. The HJ-biplot representation matrix of soil samples and studied variables.

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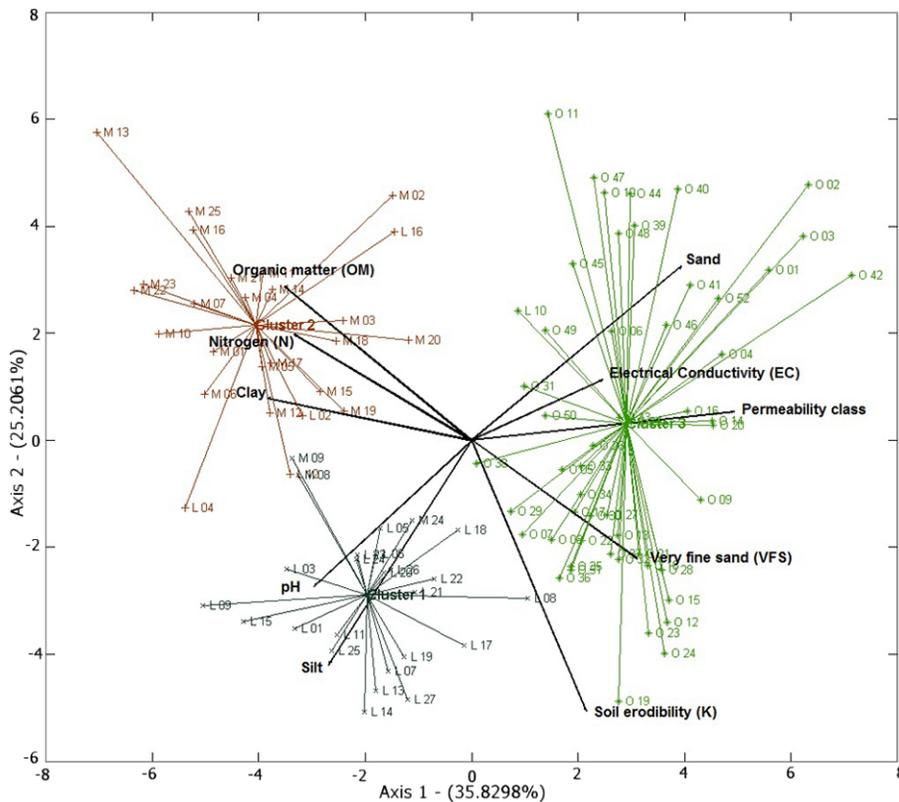


Figure 5. Hierarchical clusters representation of soil samples and studied variables.

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