

Effects of Vegetation Restoration on the Aggregate Stability and Distribution of Aggregate-Associated Organic Carbon in a Typical Karst Gorge Region

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Abstract

Land use changes have a major impact on soil structure and soil nutrients. The influences of vegetation restoration on aggregate stability and soil carbon storage have been studied extensively, but the distribution of aggregate-associated carbon is not yet understood. The objective of this study was to discuss the influences of vegetation restoration on aggregate stability and distribution of soil organic carbon (SOC) associated in water-stable aggregates (WSA) in karst gorge region. The experiment was carried out in 2012, including four vegetation types: bare land (BL), grassland (GL), shrubland (SL), and woodland (WL). Soil samples were collected from the 0-20, 20-40, and 40-60cm depths, and aggregates were separated by wet-sieving method. Aggregate stability and aggregated-associated SOC were determined, and the relationships between water-stable aggregation with SOC were examined. The results showed that total SOC and SOC associated in WSA of various sizes were the highest at a soil depth of 0-20 cm. In addition, the SOC contents of the WSA increased as the soil aggregate sizes decreased. The SOC contents of the WSA <0.25 mm were highest except in the bare land, and the SOC contents of the aggregates <0.25 mm comprised the majority of the total aggregate SOC contents. The aggregates were dominated by particles with sizes >5 mm under dry sieving treatment, while aggregates were predominantly comprised of WSA

删除的内容: Changes in soil utilization significantly affect aggregate stability and aggregate-associated soil organic carbon (SOC). A field investigation and indoor analysis were conducted in order to study the soil aggregate stability and organic carbon distribution in the water-stable aggregates (WSA) of the bare land (BL), grassland (GL), shrubland (SL), and woodland (WL) in a typical karst gorge region. The results indicated that the BL, GL, SL, and WL

删除的内容: and that the soil aggregate contents of various sizes decreased as the particle size decreased. In addition, the BL, GL, SL, and WL

1 <0.25 mm under wet sieving treatment. At a soil depth of 0-60 cm, the mean weight diameter
2 (MWD), geometrical mean diameter (GMD), and fractal dimensions (D) of the dry aggregates and
3 water-stable aggregates in the different types of land were ranked, in descending order, as
4 WL>GL>SL>BL. The contents of WSA >0.25 mm, MWD and GMD increased significantly, in that
5 order, and the percentage of aggregate destruction (PAD) and fractal dimensions decreased
6 significantly as the soil aggregate stability improved. SOC contents increased after vegetation
7 restoration, and the average SOC content of WL was 2.35 times, 1.37 times, and 1.26 times greater
8 than that in the BL, GL, and SL, respectively. We conclude that woodland and grassland facilitated
9 WSA stability and SOC protection, thus, promoting the natural restoration of vegetation by
10 reducing artificial disturbances could effectively restore the ecology and prevent soil erosion in
11 karst regions.

12 **Keywords:** Soil aggregates; Stability; Soil organic carbon; Vegetation restoration; Karst gorge

14 1 Introduction

15 Soil aggregates are the basic units of soil structures, and act as carrier of soil carbon
16 sequestration function and carbon stabilization (Cerdà 1996; Keesstra, 2012; Brevik, 2015). Good
17 soil structures provide solid foundations for the storage and stabilization of organic carbon (Jastrow,
18 1996; Mao et al., 2007; Gelaw, 2013). The particle sizes of aggregates affect their abilities to store
19 organic carbon as well as the distributions of their stored organic carbon components (Abu-Hamdeh
20 et al., 2005; Liu et al., 2009). The distribution and stability of soil aggregates are closely related to
21 the erosion resistance of soil and, thereby, are effective indicators of erosion sensitivity (Guo et al.,
22 2007; Rachman et al., 2003; Valmis et al., 2005). The results of studies conducted by Le Bissonnais
23 (1996; 1997) indicated that soil erosion primarily result from the destruction of soil aggregates.
24 Young (1980) and Bryan (2000) determined that aggregate stability affects the erodibility and
25 nutrient holding capacity of soil. The formation and stability of the water-stable aggregates in soil
26 are dependent on soil organic carbon (SOC), simultaneously vegetation communities affect soil
27 organic carbon content via the addition of outer soil organic matter and in turn contribute to the
28 formation of soil aggregates (Gabarrón-Galeote, 2013; Mekonnen, 2015). Impact of land use
29 changes on aggregate stability and distribution of aggregate-associated SOC have always been
30 research hotspots (Unger et al., 1997; Dimoyiannis et al., 2012; Stanchi et al., 2015). Jastrow (1996)

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1 researched the formation and stabilization of macroaggregates and process of C aggradation under
2 different disturbances. Burri (2009) concluded that revegetation measures increased soil aggregate
3 stability by substantially accelerating vegetation development and by promoting soil formation
4 process. Also, Mataix-Solera (2011) found that low severity fires do not produce notable changes in
5 aggregate stability, but high severity fires can induce important changes in this property.

6 China's karst region comprises an area of $3.44 \times 10^6 \text{ km}^2$. The ecosystems that have developed
7 from the karst landforms in this region are characterized by simple ecological community structures,
8 small environmental capacities, and weak resistance to interference. These ecologically vulnerable
9 areas typical to China are subjected to significant land degradation and stony desertification (Yuan
10 et al., 2002; Yan et al., 2015). The SOC pool is the largest carbon pool in the karst system (Pan and
11 Cao, 1999); the transfer of carbon in the karst system is predominantly controlled by soil carbon.
12 Excessive land utilization and management often result in the destruction of soil structures, the
13 disturbance of the foundations for organic carbon sequestration, the acceleration of soil carbon pool
14 activity, and increased levels of soil erosion (Bai et al., 2013; Tang et al., 2014). Previous studies
15 concerning the stability of soil aggregates and the characteristics of the organic carbon in those
16 aggregates have primarily been conducted in the Loess Plateau (Liu et al., 2013; Qi et al., 2011) and
17 hilly red soil regions (Guo et al., 2007; Yan et al., 2007). Studies regarding the karst region have
18 only recently been conducted. Due to strong karstification, the karst region possesses a unique
19 surface-underground structure and soil erosion different from that in the Loess Plateau and hilly red
20 soil regions, showing that soil leakage underground occurs in addition to soil erosion on the surface
21 (Zhang et al., 2011; Xu, 2014; Yan et al., 2015). Previous studies concerning the distribution of the
22 soil aggregates, aggregate stability, and distribution and mineralization of the organic carbon in the
23 aggregates (Wei et al., 2011; Tan et al., 2014) of the karst region have primarily consisted of
24 single-factor studies. In addition, due to the differences in surface vegetation, litter, and roots
25 resulting from the strong spatial heterogeneity of karst soil, the properties of the soil in different
26 regions varies greatly (Li et al., 2013). Soil degradation in karst region is characterized by
27 imbalance of soil aggregates constituting and stability decrease. Soil erosion and tillage tend to
28 damage WSA, and fine particles as well as SOC wrapped inside are susceptible to surface water
29 migration loss. Thus, further investigation on changes of SOC content and aggregates stability in
30 process of vegetation degradation/restoration has important theoretical and practical significance in

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1 revealing the evolution of soil quality in karst region.

2 In this paper, the effects of vegetation restoration on aggregates stability and the distribution as
3 well as accumulation of aggregate-associated SOC were analyzed by studying grassland (GL),
4 shrubland (SL), woodland (WL), and bare land (BL) soils typical to the gorge region of the karst
5 plateau in Guizhou Province, China. Furthermore, the influencing mechanism of vegetation
6 restoration on the stability of soil structures and sequestration of organic carbon were investigated
7 in order to provide a scientific basis for future studies regarding the changes in the carbon
8 source/sink functions of the soil in the karst region and provide a reference concerning the
9 restoration and reconstruction of degenerated karst ecosystems.

10 2 Materials and Methods

11 2.1 Study area

12 The study area was located in the Huajiang Gorge (25° 40' ~25° 42' N, 105° 37' ~105° 39'
13 E) demonstration area of Guanling County in Guizhou Province, China. This area, located on the
14 eastern slope of the Yunnan-Guizhou Plateau tilting toward the hills in Guangxi, is a typical gorge
15 region on the karst plateau, with an altitude of 500-1200 m and a relative height difference of 700 m.
16 This region is characterized by a mid-subtropical humid monsoon climate, with sufficient heat, an
17 annual average temperature of approximately 18 °C, and an average annual rainfall of 1200 mm.
18 The typical soils in this area are Calcaric Leptosols according to WRB-based soil classification,
19 which are badly structured, dry and barren.

20 The zonal vegetation in this area is comprised of mid-subtropical broadleaved evergreen
21 forests. Due to the influence of several factors, such as lithology, drought, soil, and human activity,
22 this area has experienced significant levels of vegetative degradation and is characterized by fragile
23 ecosystems and a small environmental carrying capacity. The arbor forests, shrubs, and herbs in this
24 region primarily consist of *Pteroceltis tatarinowii*, *Tona sinensis*, and *Sapium sebiferum*;
25 *Pyracantha fortuneana*, *Dodonaea viscosa*, *Zanthoxylum bungeanum*, and *Rosa cymosa*; and
26 *Imperata cylindrica*, *Arthraxon hispidus*, *Taraxacum mongolicum*, and *Dicranopteris dichotoma*,
27 respectively. Four typical vegetation types including: bare land (BL), grassland (GL), shrubland (SL)
28 and woodland (WL) were selected for studying the changes of aggregates stability and
29 aggregate-associated SOC in June 2012. The land types were selected based on the topography unit

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1 features and vegetation restoration status, in accordance with the principle of typical and
2 representative. It should be noted that bare land, grassland, shrubland and woodland were farmlands
3 before vegetation restoration. The characteristics of the sample plots are shown in Tab. 1.

删除的内容: Four types of vegetation typical to this area, including bare land (BL), grassland (GL), shrubland (SL) and woodland (WL)

4 2.2 Sample collection and analysis

5 Three research plots respectively with a 20m×20m horizontal projection area were established
6 for each vegetation type. After removing litters from soil surface, undisturbed soil were sampled at
7 depths of 0-20cm, 20-40cm and 40-60cm by quincunxes method using a shovel in each research
8 plot. The soil samples of each soil layer from the five points were mixed together as the plot's
9 samples, and a total of thirty-six mixed soil samples were collected for the purposes of the study.
10 All soil samples were brought back to laboratory and spread flat on kraft papers, then broken into
11 10-mm clods along their soil cracks, and air-dried indoors.

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12 The separation and water stability of aggregates was determined using the conventional
13 dry-sieving and wet-sieving method (Perfect et al., 1992; Unger et al., 1997). This method was
14 used to identify the disintegration processes of the soil aggregates under dissipation and expansion
15 (Qi et al., 2011). Firstly, the air-dried soil samples were mixed well, and approximately 1 kg of soil
16 samples was obtained by quartering and sieving the samples with sieves with mesh sizes of 8 mm, 5
17 mm, 2 mm, 1 mm, 0.5 mm, and 0.25 mm. Secondly, a total of 100 g of the dry-sieved aggregates
18 with different size fractions was weighed and placed on sieves with mesh sizes of 5 mm, 2 mm, 1
19 mm, 0.5 mm, and 0.25 mm. The soil samples became saturated after being wet for 10 minutes. The
20 soil samples were shaken vertically for five minutes at an amplitude of 3 cm and a frequency of 30
21 times/minute. Finally, the residue on the sieves was collected, dried at 60 °C, and weighed (Sainju
22 et al., 2003). The total SOC and SOC associated in water-stable aggregates were determined
23 through oxidation with potassium dichromate and external heating (Bao, 2005).

24 2.3 Data calculation and analysis

25 The aggregate stability index (ASI) was determined by transfer matrix method (Shi, 2005).
26 The percentage of aggregates destruction (PAD), mean weight diameter (MWD) and geometrical
27 mean diameter (GMD) were determined by dry-wet sieving method (Yan et al., 2007; He et al.,
28 2011). Since fractal theory can characterize the soil particle distribution and structural features, it has

1 been widely applied to the study of soil structure fractal since the 1980s. Fractal dimension based on the weight
2 distribution were calculated to characterize the distribution and stability of aggregates (Tyler and
3 Wheatcraft, 1989; Yang et al., 1993).

4 All statistical analyses were performed using Excel 2003 and SPSS 18.0 software. The data
5 conforms to normal distribution upon examination. The one-way analysis of variance (one-way
6 ANOVA) and least significant difference (LSD) values were used to compare the differences among
7 the various data sets. Pearson's correlation coefficient was used to evaluate the correlations among
8 the different factors. The significance level was defined as $\alpha=0.05$.

9 3 Results

10 3.1 Effects of vegetation restoration on total organic carbon and distribution of 11 SOC associated in water-stable aggregates

12 The SOC contents of soil aggregates with various particle sizes can be used to
13 micro-characterize the balance between organic matter and the mineralization rate of organic carbon.
14 Thus, these contents significantly affect the soil fertility and carbon sinks of soil. As shown in Fig. 1,
15 the organic carbon contents of the aggregates with various particle sizes differed significantly based
16 on the type of vegetation. The total organic carbon content of the study area ranged from 10.25
17 $\text{g}\cdot\text{kg}^{-1}$ to 34.07 $\text{g}\cdot\text{kg}^{-1}$. The organic carbon contents of the soil in the WL, SL, GL, and BL were
18 highest at a soil depth of 0-20 cm and decreased as the soil depth increased. The total organic
19 carbon contents of the various soil layers were ranked, in descending order, as $\text{WL} > \text{SL} > \text{GL} > \text{BL}$.

20 The contents of SOC associated in water-stable aggregates various between different size
21 fractions. In the BL, at a soil depth of 0-40 cm, the organic carbon contents of the aggregates with
22 sizes of 0.5-1 mm and 0.25-0.5 mm were the highest; at a soil depth of 40-60 cm, the organic
23 carbon content of the aggregates <0.25 mm was highest. However, this difference was not
24 significant ($P>0.05$). In the GL, SL, and WL, the contents of the aggregates <0.25 mm were the
25 highest. In general, the organic carbon content decreased as the particle size increased. In addition,
26 the organic carbon content of the aggregates >5 mm was the lowest and differed significantly from
27 those of the aggregates with different particle sizes, indicating that the organic carbon in the WSA
28 with larger particle sizes exhibited more sensitive responses to forest restoration.

29 The organic carbon contents of the soil aggregates with various sizes were highest in the

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$$y = x_1 \times \frac{x_2}{w} \times 100\% \dots \dots \dots (1)$$

Where: x_1 = percent of aggregates in this size fraction (%); x_2 = content of SOC associated in aggregates ($\text{g}\cdot\text{kg}^{-1}$); w = contents of total organic carbon in soil ($\text{g}\cdot\text{kg}^{-1}$). .

The transfer matrix method proposed by Shi (2005) was used to evaluate the aggregate stability index (ASI). The basic equations are: .

$$M_i X_i = N_i \dots \dots \dots (2)$$

$$\text{ASI} = X_1 + X_2 + X_3 + \dots + X_i \dots \dots \dots (3)$$

Where: M_i = matrix of dry-sieved aggregates contents in i particle size ranges; N_i = matrix of wet-sieved aggregates contents in i particle size ranges; X_i = the probability that each aggregate size would remain unchanged. .

The percentage of aggregate destruction (PAD) was expressed as: .

$$\text{PAD} = \frac{w_1 - w_2}{w_1} \times 100\% \dots \dots \dots (4)$$

Where: w_1 = contents of dry-sieved aggregates >0.25 mm (%); w_2 = contents of water-stable aggregates >0.25 mm (%). .

The mean weight diameter (MMD) and geometrical mean diameter (GMD)

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1 topsoil. In the BL, the organic carbon contents of the aggregates with particle sizes of >5 mm, 1-2
2 mm, and 0.25-0.5 mm initially decreased, then increased as the soil depth increased, and the organic
3 carbon contents of the aggregates with other size fractions decreased as the soil depth increased. In
4 the SL, the organic carbon contents of the aggregates with particle sizes of 0.25-0.5 mm and <0.25
5 mm initially decreased, then increased as the soil depth increased, and the organic carbon contents
6 of the aggregates with other particle sizes decreased as the soil depth increased. In the GL and WL,
7 the organic carbon contents of the soil aggregates with various size fractions decreased as the soil
8 depth increased.

9 As shown in Tab.2, in all of the types of vegetation, the contribution of the organic carbon in
10 the WSA<0.25 mm to the total organic carbon content of the soil was highest, ranging from 18.85%
11 and 41.08%, with an average of 25.95%. In the BL, GL, and SL, the contribution of the organic
12 carbon contents of the aggregates >5 mm was lowest with values of less than 10%. In the WL, the
13 contribution of the organic carbon contents of the WSA with sizes of 0.25-0.5 mm was the lowest.
14 At different soil depths, the contributions of the organic carbon contents of the WSA with various
15 sizes to the total organic carbon contents of the aggregates varied insignificantly.

16 **3.2 Effects of vegetation restoration on distribution of soil aggragates,**

17 As shown in Fig.2, the constituent size fractions of the dry-sieved aggregates in the different
18 types of vegetation differed. For all vegetation types, the dry-sieved aggregates in the different
19 layers of soil predominantly consisted of aggregates ≥ 2 mm, accounting for greater than 60% of
20 the total aggregates. The aggregates >5 mm also comprised a large amount of the soil aggregates,
21 accounting for 35.56-60.98% of the total aggregates. The aggregates <0.25 mm comprised the
22 smallest proportion of aggregates, accounting for 3.52-8.49% of the total aggregates. In general, the
23 aggregate contents decreased as the particle size decreased. Each size fraction of the soil aggregate
24 contents varied among the different types of vegetation, but not importantly. The soil aggregate
25 contents of different size fractions varied somewhat within each type of vegetation.

26 For all four vegetation types, the contents of the aggregates >5 mm decreased as the soil depth
27 increased, the contents of the aggregates <0.25 mm increased as the soil depth increased, and the
28 contents of the aggregates with other particle sizes varied differently as the soil depth varied. Within
29 the various layers of soil, the contents of the aggregates >5 mm were ranked, in descending order,

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1 as WL>GL>SL>BL, and the contents of the aggregates ranging from 2 to 5 mm were ranked, in
2 descending order, as SL>BL>GL>WL. The contents of the aggregates <0.25 mm in the top soil
3 decreased as the vegetation restoration progressed.

4 As shown in Fig.3, the water-stable aggregates of the different types of vegetation
5 predominantly consisted of particles with sizes <0.25 mm, 1-2 mm, and 2-5 mm, accounting for
6 59.01-71.31% of the total WSA. In general, the WSA contents initially increased, then decreased,
7 and then increased as the particle sizes decreased. The WSA <0.25 mm comprised the majority of
8 the total WSA, accounting for 23.64-35.93% of the total WSA. Within the various layers of soil, the
9 content of aggregates <0.25 mm in the different types of vegetation increased as the soil depth
10 increased according to the ranking BL>SL>GL>WL, while the contents of the larger aggregates
11 decreased as the soil depth increased. The BL exhibited the highest content of aggregates <0.25 mm
12 in the various layers of soil. The contents of aggregates <0.25 mm in the other vegetation types
13 were less than 30%.

14 3.3 Effects of vegetation restoration on soil aggregate stability

15 As shown in Table.3, the stability of the soil aggregates differed significantly based on the type
16 of vegetation. The soil depth also affected the stability of the soil aggregates. The contents of the
17 aggregates >0.25 mm in the WL, GL, SL, and BL, in descending order, were equal to 81.37%,
18 79.49%, 69.02%, and 68.65%, respectively. The PAD of the aggregates >0.25 mm in the BL
19 (27.24%) was the highest, and the PAD of the aggregates >0.25 mm in the WL (16.27%) was the
20 lowest. The MWD and GMD values of the different types of vegetation varied consistently. The
21 MWD and GMD values of the dry-sieving and wet-sieving aggregates were both ranked, in
22 descending order, as WL>GL>SL>BL. The fractal dimensions of the dry-sieved aggregates in the
23 four types of vegetation ranged from 1.994 to 2.227, and the fractal dimensions of the WSA ranged
24 from 2.425 to 2.725, ranking, in descending order, as BL>SL>GL>WL at a soil depth of 0-60 cm.

25 In all four types of vegetation, the contents of the aggregates >0.25 mm, MWD and GMD,
26 decreased as the soil depth increased. In addition, the PAD and D values increased as the soil depth
27 increased. The MWD and GMD of the dry-sieved and wet-sieved aggregates in the various layers of
28 soil were ranked, in descending order, as WL>GL>SL>BL, while the D values of the dry-sieving
29 and wet-sieving aggregates in the various layers of soil were ranked, in descending order, as

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1 BL>SL>GL>WL. The PAD₁ of the aggregates obtained at a soil depth of 0-20 cm were ranked, in
2 descending order, as SL>BL>GL>WL, while the PAD₂ of the aggregates obtained at soil depths of
3 20-40 cm and 40-60 cm were ranked, in descending order, as BL>SL>WL>GL. The stability of the
4 soil aggregates in the same soil layers of the different types of vegetation varied significantly
5 ($P<0.05$). Other indicators changed not significantly as the soil depth changed.

6 The probabilities of the soil aggregates in the different types of vegetation remaining
7 unchanged are shown in Table.3. In the BL, GL, and SL, dry-sieving and wet-sieving, essentially
8 destroyed the aggregates with particle sizes greater than 5 mm and ranging from 1 mm to 5 mm,
9 and somewhat affected the aggregates with sizes ranging from 0.25 mm to 1 mm. The probability of
10 the soil aggregates with sizes ranging from 0.25 mm to 1 mm remaining unchanged ranged from
11 0.26 and 0.51. In the WL soil, the probability of the soil aggregates >5 mm remaining unchanged
12 was the highest (0.39-0.55), and the probability of the soil aggregates with sizes ranging from 0.25
13 mm to 0.5 mm only ranged from 0.23 to 0.31, indicating that these particles were easily broken
14 when subjected to dissipation and disintegration. The ASI values varied significantly, ranging from
15 2.19 to 3.32. The average ASI values of the WL, GL, SL, and BL were, in descending order, equal
16 to 2.85, 2.65, 2.39, and 2.31, respectively. The ASI values of the aggregates in the topsoil were the
17 highest, and the ASI values decreased as the soil depth increased. The ASI values of the different
18 types of vegetation at various soil depths were ranked, in descending order, as WL>GL>SL>BL.
19 The differences in the ASI values of the GL and WL, as well as BL and SL at the same soil depths
20 were not significant. However, except at soil depths of 20-40 cm and 40-60 cm, the differences in
21 the ASI values of the GL and BL and the WL and SL at the same soil depths were significant.

22 3.4 Relationships between water stable aggregates and organic carbon

23 The correlations among the parameters of the WSA are shown in Table 5. D was significantly
24 and negatively associated with the MWD, GMD, and SOC ($P<0.01$), and the MWD was
25 significantly and positively associated with the GMD ($P<0.01$). The SOC was significantly and
26 positively associated with both the MWD and GMD ($P<0.01$), indicating that, as the SOC increased,
27 the MWD and GMD also increased, promoting the stability of the soil structures. The contents of
28 the aggregates with particle sizes of greater than 5 mm, 2-5 mm, and 1-2 mm were significantly and
29 negatively correlated with D ($P<0.05$), and the contents of the aggregates with particle sizes of

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The SOC contents of soil aggregates with various particle sizes can be used to micro-characterize the balance between organic matter and the mineralization rate of organic carbon. Thus, these contents significantly affect the soil fertility and carbon sinks of soil. As shown in Fig.3, the organic carbon contents of the aggregates with various particle sizes differed significantly based on the type of vegetation. The total organic carbon content of the study area ranged from $10.25 \text{ g}\cdot\text{kg}^{-1}$ to $34.07 \text{ g}\cdot\text{kg}^{-1}$. The organic carbon contents of the soil in the WL, SL, GL, and BL were highest at a soil depth of 0-20 cm and decreased as the soil depth increased. The total organic carbon contents of the various soil layers were ranked, in descending order, as WL>SL>GL>BL. The organic carbon contents of the aggregates with different size fractions differed, though not significantly, based on the type of vegetation. In the BL, at a soil depth of 0-40 cm, the organic carbon contents of the aggregates with sizes of 0.5-1 mm and 0.25-0.5 mm were the highest; at a soil depth of 40-60 cm, the organic carbon content of the aggregates <0.25 mm was highest. However, this difference was not significant ($P>0.05$). In the GL, SL, and WL, the contents of the aggregates <0.25 mm were the highest. In general, the organic carbon content decreased as the particle size increased. In addition, the organic carbon content

0.25-0.5 mm and less than 0.25 mm were significantly and positively correlated with D ($P<0.05$). The MWD and GMD were significantly and positively correlated with the contents of the aggregates >2 mm and significantly and negatively correlated with the contents of the aggregates <0.5 mm. The SOC was positively correlated with the contents of aggregates with various size fractions and significantly correlated with the contents of aggregates greater than 5 mm, indicating that, as the content of soil aggregates with large size fractions increased, the stability of the soil structures and the levels of soil organic carbon increased.

4 Discussion

4.1 Effects of vegetation restoration types on distribution of SOC associated in water-stable aggregates

The organic carbon contents of aggregates, which reflect the balance and mineralization of organic carbon, significantly affect the nutrient holding capacity and carbon sequestration in soil (Wu et al., 2004). The humus horizons of the various types of vegetation differed based on the quantity and quality of the litter and the environment, affecting the organic carbon contents of the soil and the stability and contents of the organic carbon in the aggregates (Novara et al., 2015). In all four types of vegetation, the organic carbon contents in the aggregates with various particle sizes were the highest at a soil depth of 0-20 cm. These results were consistent with the results of a study conducted by Li et al. (2008). This was because the large amount of plant residue that had accumulated in the topsoil and the amount of organic matter that had been input into the soil improved the biological activity of the microorganisms, animals, and roots in the topsoil and, thus, facilitated the formation of particulate organic carbon (Wei et al., 2011).

In this study, as the vegetation transitioned from BL to WL, the total soil organic carbon contents and the organic carbon contents in the aggregates with various particle sizes increased significantly. The organic carbon contents of the aggregates were the highest in the WL, and the lowest in the SL and BL, primarily due to the amount of vegetative coverage and the quantity and decomposition of litter. Due to their high amounts of vegetative coverage, the WL and GL exhibited large amounts of litter and a considerable amount of input soil organic carbon. The BL and SL exhibited significantly smaller amounts of litter and input soil organic carbon and accelerated levels of organic carbon decomposition due to artificial disturbances. In general, the organic carbon

contents of the soil aggregates decreased as the soil depth increased; in these types of vegetation, artificial disturbances accelerated the decomposition of organic carbon.

In previous studies, De Jonge (1999), Christensen (1986), and Li et al. (2006) determined that organic carbon is primarily distributed in micro-aggregates (<0.25 mm) and that organic carbon contents increase as aggregate particle sizes decrease. Puget (1998; 2000) found that large aggregates are a source of organic carbon enrichment. In another study, Li et al. (2000) found that organic carbon is distributed in a "V" shape in aggregates and that the organic carbon contents of aggregates >2 mm and <0.25 mm are high. In Li's study, as the particle sizes increased, the soil organic carbon contents decreased, but the organic carbon contents of the aggregates exhibited no significant differences, possibly due to the high calcium carbonate and clay contents of the lime soil in the karst region (Wei et al., 2011). In this study, the organic carbon contents of the aggregates <0.25 mm in the grassland, shrubland, and woodland increased by 5.28%-95.37%, 1.46%-106.25%, and 6.02%-85.43% compared to the aggregates with other particle sizes. These results corresponded with the theory that organic carbon initially accumulates in aggregates with small size fractions (Hassink, 1997) as well as the results of other studies concerning the karst region (Lu et al., 2012; Luo et al., 2011).

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4.2 Effects of vegetation restoration on the distribution and stability of soil aggregates

The size fractions of aggregates affect the storage and supply of soil nutrients, the pore structure and hydraulic properties of soil, and the movement of organisms in soil. Therefore, the size distributions of aggregates are closely related to soil quality (Dexter 1988; Nimmo and Perkins, 2002). Soil organic carbon is a binding substance imperative for the formation of aggregate structures (Cerdà 2000; Wu et al., 2004). Land utilization significantly affects soil organic carbon contents by influencing the input and output of the organic matter in soil and, thereby, the distribution and stability of soil aggregates (Power 2004; Luo et al., 2011; Sajjadi, 2014). After vegetation restoration, the organic carbon contents, MWD, and GMD of the WSA >0.25mm increased, and the PAD and D values decreased. Thus, as a result of vegetation restoration, the soil organic carbon content increased, promoting the formation of soil aggregates and increasing the stability of the soil aggregates. The organic carbon contents, MWD, and GMD of the

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WSA >0.25mm in the WL and GL were significantly higher than those of the WSA >0.25mm in the BL and SL, while the PAD values of the WL and GL were significantly lower than those of the BL and SL, indicating that the WL and GL possessed good soil aggregate structures with strong corrosion resistance and high nutrient storage capacities. In the BL and SL, artificial disturbances led to the destruction of soil structures, accelerating the transformation of large WSA to aggregates with small size fractions and exacerbating the low of soil and nutrients. The structural stability and erosion resistance of the soil in the GL were higher than those of the soil in the SL, primarily because the GL was subjected to less artificial disturbances. Thus, the GL exhibited higher levels of vegetative coverage with considerable amounts of returned biomass, while the SL, due to felling, picking, and the forced compaction of wildlife, possessed fewer soil organic matter sources and disrupted soil structures. Bare land and shrubland were prone to erosion when suffer strong rainfall splash due to low surface vegetation coverage, leading to aggregates dispersed into small ones and decomposition of unstable organic carbon, which result in a substantial decline in macro-aggregates and organic carbon content (Gabarrón-Galeote, 2013).

The contents of the WSA decreased as the soil depth increased. This was likely because the excess litter on the topsoil; the high organic matter contents; and the good water, heat, and air conditions of the soil contributed to the formation of large aggregates (Tisdall and Oades, 1982; Xiao et al., 2008). In addition, the organic matter contents of the deep soil were relatively low, a small number of large soil aggregates were formed, and the contents of the large WSA were reduced. In the same types of vegetation, the ASI of the WSA decreased as the soil depth increased, just as the contents of the large WSA decreased as the soil depth increased. Of the four typical types of vegetation in the study area, the WL exhibited the highest aggregate stability and strongest erosion resistance, promoting the stability of the soil structures, the storage of nutrients, and the contents of organic carbon. The grassland exhibited the second highest aggregate stability, and the bare land and shrubland possessed relatively poor water stability.

The vegetation restoration process significantly affected the formation and distribution of large aggregates in that the BL, SL, GL, and WL exhibited significantly decreased levels of aggregates <0.25 mm throughout the various soil layers and significantly increased levels of larger aggregates throughout the vegetation restoration process. Thus, vegetation restoration effectively improved the soil infiltration capacity, water holding capacity, and aeration of the lime soils. Therefore, aggregate

1 stability could be improved and soil erosion could be prevented by reducing artificial disturbances,
2 increasing the organic matter and nutrient contents of soil, and, thereby, facilitating the natural
3 restoration of vegetation.

4 The fractal dimensions of the particle size distributions of soil granular structures reflect the
5 influence of the contents of aggregates on the structure and stability of the soil (Dexter 1988;
6 Nimmo and Perkins, 2002). Thus, smaller fractal dimensions indicate better soil structures and
7 stability and higher erosion resistance. Likewise, higher fractal dimensions indicate poorer soil
8 structures and stability (Zhou et al., 2008; Barral et al.,1998).

9 The fractal dimensions of various soil particle sizes reflect the ability of the soil particles to fill
10 spaces and could be used to evaluate soil structures (Tyler and Wheatcraft, 1989; Yang et al., 1993).
11 D was significantly and negatively correlated with the MWD, GMD, and SOC, indicating that, as
12 the value of D increased, the contents of the aggregates >0.25 mm decreased, and the soil density
13 increased, resulting in poorer permeability and nutrient and moisture storage capacities. The results
14 of this study indicated that, as the vegetation was restored, the fractal dimensions of the WSA
15 decreased, and the stability of the soil structures and organic carbon contents improved.

16 **5 Conclusions**

17 Total SOC and SOC associated in water-stable aggregates increased with the restoration of
18 vegetation. SOC was initially accumulated in aggregates with small size fractions and
19 WSA<0.25mm has the greatest contribution to soil total SOC. Contents of WSA<0.25mm reduced
20 while WSA>1mm increased significantly in the process of vegetation succession from bare land to
21 woodland, vegetation restoration has promoted the accumulation of aggregates with small particle
22 sizes into large sizes and, thereby, improved water stability of aggregates. The fractal dimension of
23 water stable aggregates was highly significantly and negatively correlated with the MWD, GMD,
24 and SOC (P<0.01), indicating that it could be used to objectively and comprehensively reflect the
25 soil aggregate characteristics and stability. The woodland and grassland were more conducive to
26 facilitate WSA stability and SOC protection, thus, promoting natural vegetation restoration by
27 reducing artificial disturbances could effectively restore the ecology and prevent soil erosion in
28 karst regions.

29 **Acknowledgments**

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<#>Fractal characteristics of the aggregates and their responses to vegetation restoration .
The size fractions of aggregates affect the storage and supply of soil nutrients, the pore structure and hydraulic properties of soil, and the movement of organisms in soil. Organic carbon is an important material for the formation of soil aggregates, vegetation communities affect soil organic carbon content via the addition of outer soil organic matter and in turn contribute to the formation of soil aggregates (Gabarrón-Galeote, 2013; Mekonnen, 2015). The fractal dimensions of the particle size distributions of soil granular structures reflect the influence of the contents of aggregates on the structure and stability of the soil (Dexter 1988; Nimmo and Perkins, 2002). Thus, smaller fractal dimensions indicate better soil structures and stability and higher erosion resistance. Likewise, higher fractal dimensions indicate poorer soil structures and stability (Zhou et al., 2008; Barral et al.,1998).

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8
9
10 **Table 1.** Basic properties of study plots

Vegetation types	Altitude (m ^{a.s.l})	Slope (°)	Vegetation cover (%)	Dominant species	Land use
BL	696	20.2	<10	<i>Imperata cylindrica</i> , <i>Arthraxon hispidus</i> .	Farmland abandoned one year, with disturbance of tillage and pasture.
GL	710	22.1	70	<i>Imperata cylindrica</i> , <i>Leucas mollissima</i> and <i>Taraxacum mongolicum</i> .	Natural recovery for 5 years, with less human disturbance.
SL	694	25.4	60	<i>Pyracantha fortuneana</i> , <i>Rosa cymosa</i> and <i>Dodonaea viscosa</i> ..	Natural recovery for 9 years, with disturbance of pasture.
WL	704	20.0	80	<i>Tona sinensis</i> , <i>Pteroceltis tatarinowii</i> and <i>Sapium sebiferum</i> .	Natural recovery for 16years, with less human disturbance.

11
12 **Table 2.** Contribution rates of water stable aggregates organic carbon to SOC under different
13 vegetation types (% , mean \pm SE, n=3)

Vegetation types	Layer /cm	WSA sizes (%)					
		>5mm	5~2mm	2~1mm	1~0.5mm	0.5~0.25mm	<0.25mm
BL	0-20	5.38 \pm 0.33d	17.68 \pm 0.9bc	11.6 \pm 1.27c	14.82 \pm 1.08c	21.19 \pm 0.19b	25.32 \pm 0.46a
	20-40	4.92 \pm 0.19d	13.92 \pm 0.31c	12.65 \pm 0.21c	16.42 \pm 0.95b	16.78 \pm 0.16b	27.88 \pm 1.43a
	40-60	3.45 \pm 0.20e	15.87 \pm 1.05bc	13.6 \pm 0.61c	10.49 \pm 0.56d	19.78 \pm 0.96b	41.08 \pm 1.81a
GL	0-20	7.56 \pm 0.21b	17.65 \pm 0.57a	16.99 \pm 0.38a	16.44 \pm 0.45a	8.55 \pm 0.23b	18.85 \pm 0.35a
	20-40	2.43 \pm 0.07d	18.45 \pm 0.61b	22.37 \pm 0.56a	18.57 \pm 0.60b	11.26 \pm 0.31c	24.10 \pm 0.50a
	40-60	2.58 \pm 0.15d	19.3 \pm 0.95b	17.04 \pm 0.6b	17.54 \pm 0.57b	11.27 \pm 0.34c	27.72 \pm 0.73a
SL	0-20	4.24 \pm 0.11d	9.84 \pm 0.29c	9.28 \pm 0.22c	9.11 \pm 0.3c	11.71 \pm 0.15b	26.51 \pm 0.33a
	20-40	1.68 \pm 0.10d	6.56 \pm 0.41c	9.00 \pm 0.34b	9.56 \pm 0.46b	7.49 \pm 0.40c	18.28 \pm 0.65a
	40-60	1.26 \pm 0.05d	7.85 \pm 0.56c	11.26 \pm 0.42b	12.44 \pm 0.67b	11.42 \pm 0.44b	32.18 \pm 1.06a
WL	0-20	16.81 \pm 0.19bc	17.29 \pm 0.73b	14.66 \pm 0.19c	9.39 \pm 0.16d	5.49 \pm 0.62e	22.19 \pm 0.20a
	20-40	16.58 \pm 0.19b	24.38 \pm 0.78a	11.13 \pm 0.21c	7.78 \pm 0.54d	5.06 \pm 0.11d	20.37 \pm 0.36a
	40-60	13.91 \pm 0.18b	24.99 \pm 0.82a	16.25 \pm 0.1b	9.95 \pm 0.40c	6.43 \pm 0.27d	26.91 \pm 1.21a

14 Note: different small letters in the same row showed significant difference at 0.05 level among different sizes.
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1

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Table 3. The soil aggregate stability based on Саввинов method

Vegetation types	Layer /cm	WSA _{>0.25} (%)	PAD (%)	MWD (mm)		GMD (mm)		D	
				Dry	Wet	Dry	Wet	Dry	Wet
BL	0-20	71.68Ab	21.5Cb	4.183Ad	1.588Ac	3.181Ac	0.699Ac	2.163Aa	2.649Ba
	20-40	68.21ABb	28.34Ba	4.085Ad	1.304ABc	2.627Bd	0.626Bc	2.192Aa	2.690ABa
	40-60	64.07Bb	31.87Aa	3.243Bd	1.169Bc	1.966Cd	0.535Cc	2.227Aa	2.725Aa
GL	0-20	83.36Aa	13.59Cc	5.335Ab	1.966Ab	3.882Ab	1.061Ab	2.037Aa	2.487Bc
	20-40	80.81Ba	16.33Bb	5.011ABb	1.642Bb	3.605ABb	0.913Ab	2.091Aa	2.505ABc
	40-60	74.31Ca	19.43Ac	4.883Bb	1.498Bb	3.384Bb	0.881Ab	2.121Aa	2.580Ac
SL	0-20	73.86Ab	23.13Ca	4.655Ac	1.697Ac	3.35Ac	0.71Ac	2.112Aa	2.603Ab
	20-40	68.69Bb	26.87Ba	4.467Ac	1.473ABbc	3.107ABc	0.638ABc	2.187Aa	2.623Ab
	40-60	66.5Bb	28.26Ab	4.415Ac	1.252Bc	2.842Bc	0.588Bc	2.224Aa	2.678Ab
WL	0-20	86.72Aa	10.68Bd	6.101Aa	3.618Aa	4.934Aa	1.981Aa	1.994Aa	2.425Bc
	20-40	81.03Ba	17.02Bb	5.882Aa	3.027Ba	4.534Ba	1.503Ba	2.045Aa	2.522ABc
	40-60	76.36Ca	21.1Ac	5.41Ba	2.505Ba	4.087Ca	1.151Ca	2.074Aa	2.556Ac

3 Note: different small letters in the same column meant significant differences in same layer of different vegetation
4 types at 0.05 level, different capital letters in the same column meant significant differences in different soil layer
5 of same vegetation types at 0.05 level, the same in the table 3.

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Table 4. Conservation ratio of aggregates and aggregate stability index

Vegetation types	Layer /cm	Conservation ratio of aggregates (%)						aggregate stability index (ASI)
		>5mm	5-2mm	2-1mm	1-0.5mm	0.5-0.25mm	<0.25mm	
BL	0-20	0.22	0.25	0.28	0.3	0.42	1	2.47Ab
	20-40	0.15	0.28	0.23	0.28	0.38	1	2.30Bb
	40-60	0.10	0.23	0.21	0.26	0.39	1	2.19Bb
GL	0-20	0.18	0.34	0.44	0.46	0.49	1	2.91Aa
	20-40	0.05	0.24	0.38	0.42	0.51	1	2.60Ba
	40-60	0.07	0.31	0.32	0.36	0.39	1	2.45Ba
SL	0-20	0.19	0.28	0.26	0.34	0.44	1	2.51Ab
	20-40	0.09	0.21	0.25	0.34	0.45	1	2.34Bb
	40-60	0.07	0.22	0.22	0.35	0.46	1	2.32Bb
WL	0-20	0.55	0.51	0.52	0.43	0.31	1	3.32Aa
	20-40	0.48	0.35	0.30	0.31	0.25	1	2.69Ba
	40-60	0.39	0.23	0.38	0.30	0.23	1	2.53Ba

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Table 5. Correlation between parameters of water-stable aggregation

	D	MWD	GMD	SOC	WSA sizes (mm)					
					>5	5~2	2~1	1~0.5	0.5~0.25	<0.25
D	1									
MWD	-0.544**	1								
GMD	-0.608**	0.963**	1							
SOC	-0.454**	0.701**	0.756**	1						
WSA	>5	-0.203	0.409*	0.265	0.588**	1				
sizes	5~2	-0.346*	0.168	0.203	0.348*	0.798**	1			

(mm)	2~1	-0.202	0.224	0.121	0.418*	0.871**	0.808**	1			
	1~0.5	-0.215	0.194	0.172	0.410*	0.882**	0.817**	0.951**	1		
	0.5~0.25	0.312*	0.114	0.237	0.338*	0.822**	0.678**	0.947**	0.926**	1	
	<0.25	0.633**	0.074	0.112	0.368*	0.865**	0.835**	0.960**	0.922**	0.951**	1

* $P<0.05$; ** $P<0.01$.

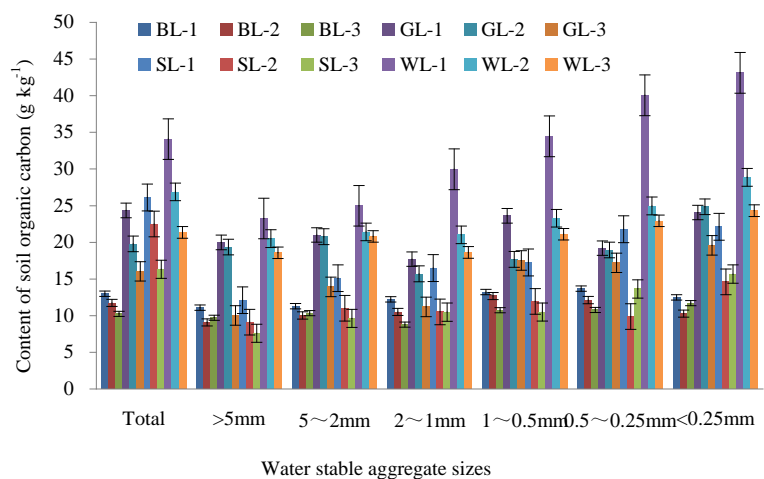


Figure 1. SOC in different water stable aggregate sizes under different vegetation types (mean ± SE, n=3)

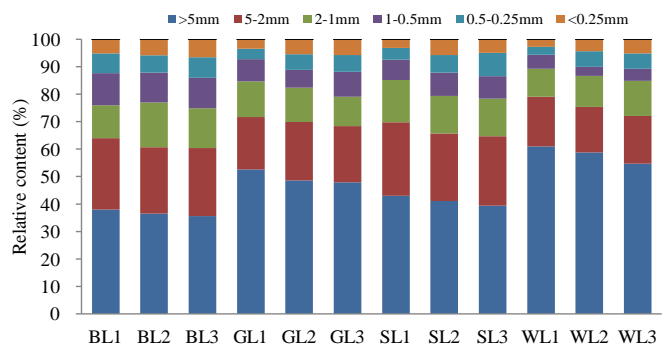
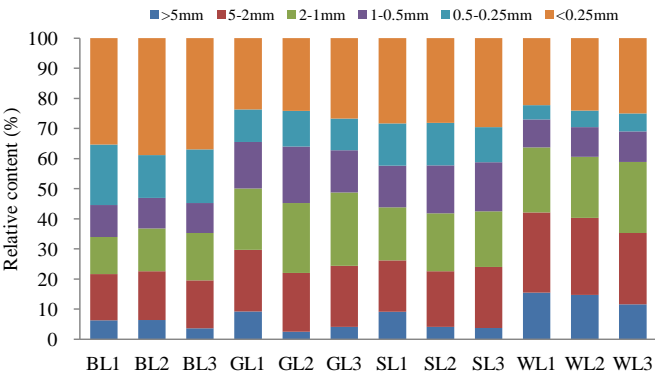


Figure 2. Relative distribution of soil dry-aggregates with different sizes under different vegetation types. (BL1, BL2, BL3, GL1, GL2, GL3, SL1, SL2, SL3, WL1, WL2, WL3 represent 0-20, 20-40 ,40-60cm soil layers for bareland, grassland, shrubland and woodland. The same in figure 2

1 and figure 3)



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3 **Figure 3.** Relative distribution of soil water stable aggregates with different sizes under different
4 vegetation types
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