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- 2 Cultivated grasslands present a higher soil organic carbon sequestration efficiency under leguminous
- 3 than under gramineous species
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12 ABSTRACT

The establishment of grasslands on abandoned cropland has been proposed as an effective method of to mitigating climate change by increasing soil organic carbon (SOC) storage. In this study, five cultivated grasslands were established (three leguminous species - Coronilla varia, Onobrychis viciaefolia, Medicagosativa, and two gramineous species-Poa annua, Agropyron cristatum), one uncultivated, and one natural grassland were studied to examine how the SOC storage, sequestration rate and sequestration efficiency to change for 5 years restoration in semi-arid area. Our results showed that the cultivated leguminous grasslands had greater total biomass, SOC storage, SOC sequestration rate and efficiency than gramineous grasslands, abandoned cropland, and natural grassland. The greater soil carbon (C) accumulation in leguminous grassland was mainly attributed to the capacity to incorporating carbon and the higher biomass production. Leguminous grasslands accumulated more SOC than gramineous grasslands by 0.64 Mg C-ha⁻¹yr⁻¹. The average SOC sequestration efficiency in leguminous grassland (1.00) was about 2 times greater than gramineous grassland (0.34). Root nodules in leguminous promote the symbiosis with micro-organisms, which are responsible for the decomposition of the plants, and therefore constitute the key of the transmission of the stored carbon into the soil. The results indicate that cultivated leguminous grasslands sequestered more SOC with higher SOC sequestration efficiency than cultivated gramineous grasslands in

- 28 arid and semi-arid areas. Our results provide a reference for ecological management in arid and semi-arid
- 29 <u>areas.</u>
- 30 **KEY-WORDS:** Cultivated grassland, Carbon sequestration, Gramineous, Leguminous, SOC

1 Introduction

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The sSoil is a key component of the Earth System and it contributes to services, goods, and resources to the humankind (Brevik et al., 2015). Soil stored stores more carbon (C) than the atmosphere and vegetation (K \(\tilde{c}\)hy et al., 2015; Keesstra et al., 2016). Soil organic carbon (SOC) as a key component of the global carbon cycle and its potential to sink from atmosphere carbon dioxide (CO₂) have been widely discussed in the scientific literatures throughout the world (Guo and Gifford, 2002; Lal, 2004; De Deyn et al., 2008; Deng et al., 2014a; Parras-Alc ántara et al., 2015). Thus during recent decades, massive emphasis had been given in SOC storage and sequestration on global scale. In the terrestrial ecosystem, SOC pool dynamics were affected by many factors, such as climate change (Lal, 2004; Field et al., 2007), management practices (Luo et al., 2010; Ono et al., 2015), land use etc. (Post et al., 2000; Don et al., 2011; Deng et al., 2014b; Muñoz-Rojas et al., 2015). SOC plays an extremely important role in controlling of soil fertility and cropping system productivity and sustainability (Hurisso et al., 2013; De Moraes S áet al., 2015), particularly in low-productivity arid and semiarid agro-ecosystems (Behera et al., 2015). To develop farming methods that conserve conserving SOC is therefore of a great importance (Lal, 2004). Cultivated grassland has much more advantages than natural grassland regeneration, such as accelerating vegetation restoration and improving grassland productivity. Establishing artificial grassland is one type of land uses to restoring vegetation and improveimproving SOC (Fu et al., 2010; Wu et al., 2010; Li et al., 2014; Wu et al., 2010). In grassland, atmospherecarbon was sequestrated through photosynthesis and respiration, then carbon fixing in stable SOC pool or releasing back into the atmosphere (Post et al., 2000). Therefore, studying the carbon sequestration in grassland ecosystems can help to identify the magnitude of global carbon sinks and sources (Li et al., 2014). The balance of Soil carbon pool is determined by the carbon input from leaf and root and itsmineralization in soil, and output in decomposition processes of soil organic matter by soil microbes and

respiration from plant roots (Amundson, 2001; Garcia Diaz et al., 2016). The biomass fraction resulting in SOC build up (plant residuals) was strongly affected by management practices including the selection of plant species (Don et al., 2011). Species composition had a great role in determining the aboveground-productivity (Liu et al., 2016). Over relatively long time, the proportion of the aboveground biomass enters soil as organic matter and incorporates into soil through physical and biological processes. For example, some leachates from plant material in the litter layer, root exudates, solid decomposed litter and fragmented plant structure materials (Jones and Donnelly, 2004; Novara et al., 2015). The amount of plant residuals returned to the soil directly affected the SOC (Musinguzi et al., 2015; Wasak et al., 2015), and mostly perennial plants were managed with high planting densities to produce greater biomass exports (Hobbie et al., 2007; Köchy et al., 2015).

Vegetation degradation and exponential population growth have caused massive amounts of soil and water to be lost. The Chinese government has implemented the most ambitious ecological program titled 'Grain-for-Green' Project (converting degraded, marginal land and cropland into grassland, shrubland and forest), with the objective of transforming the low-yield slope cropland into grassland, reducing soil erosion, maintaining land productivity and improving environmental quality (Fu, 1989; Liu et al., 2008;). The large scale of the project indeed enhanceed carbon sequestration capacity in China, especially in arid and semi-arid areas (Chang et al., 2011; Song et al., 2014).

Many prior studies about SOC have paid much attention to convertingsion from farmland to grassland, shrubland or forest (Fu et al., 2010; Deng et al., 2014a). The main dominant grass species used in the project are leguminous and gramineous (Jia et al., 2012; Wang et al., 2015). However, less attention has been devoted to the SOC among different plant species grasslands. In current study, we have focused on ascertaining the influence of leguminous and gramineous grasslands on SOC sequestration capacity and efficiency. Many studies had demonstrated that there is a significant and positive relationship on SOC and

nitrogen (Deng et al., 2013; Zhu et al., 2014). So we hypothesize that the leguminous grassland has the higher SOC sequestration capacity than gramineous grassland. More specifically, our objectives are: (i) to analyze the effects of SOC stock and sequestration differences of storage efficiency under different grasslands; (ii) to determine which type of cultivated grassland might better improve SOC storage in arid and semi-arid areas.

2 Material and methods

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2.1 Experimental site and design

The study-site was—was conducted at the Lanzhou scientific observation and experiment field station of ministry of agriculture for ecological system in the Loess Plateau area located in Gongjiawan County-(103 °44.342′ E, 36°02.196′ N, 1966-1635 m a.s.l.) of in Lanzhou, Gansu Province, China (Fig. 1). The site is the semi-arid temperate continental temperate monsoon climate zone. The data from the National Meteorological Information Center of China showed that the mean annual temperature was is 9.3 °C (2008-2012), and the minimum and maximum values were 23.1 ℃ and 39.1 ℃ (2008-2012), respectively. The annual cumulative temperature above 10 °C was between 1900 and 2300 °C·d, and above 0 °C it was 3700 °C·d. The mean annual precipitation was 324.5 mm, and which approximately 80% falls during the growing season (from May to September). The topography of study area was is typical characteristics of the Loess Plateau, such as plains, ridges and mounds, etc. The elevation of study site was about 1700 m. Soil parent material is the Quaternary eolian loess. The main soil type was is Sierozem, which is a calcareous soil and characteristics of the Chinese loess region (Li et al., 2010). Sierozem is the soil developed in the dry climate and desert steppe in warm temperate zone, which has low humus and weak leaching (National soil census office, 1998). There is the patch or pseudohyphae calcium carbonate deposition and strong lime reaction within full sierozem profile (Shi, 2013). Soil total nitrogen content is about 0.41g kg⁻¹, total phosphorus content is 0.46 g kg⁻¹, total potassium content is 18.24 g kg⁻¹, and pH is 8.25 in study site.

The experimental site was originally under sorghum (*Sorghum bicolor* L.) continuously from 1970 to 2005 and it was abandoned from 2005 to 2007 (grazing exclusion). In 2007, five cultivated grasslands, one uncultivated grassland (abandoned cropland, Un-G), one natural grassland (Na-G) were established in the study site. Five main forage grasses, widely grown across in semi-arid areas, were selected to establish five types of cultivated grassland, namely three leguminous species (*Coronilla varia* L., L-CV; *Onobrychis viciaefolia* Scop, L-OV; *Medicago sativa* L., L-MS) and two gramineous species (*Poa annua* L., G-PA; *Agropyron cristatum* L. Gaertn., G-AC) (Table 1). The seeding rates in different grassland were showed in Table 1. The different seeding rates were contributed to the percentage of germination, to guarantee the equal plant density in each grassland. The rates were determined by reference to local farmland crop density. We designed the experiment to be a randomized plot design. Three experimental plots [10 m × 20 m] were established randomly within each of the grassland areas. The forage grasses were planted in early April of 2007, and all plots were weeded manually and watered three times (April, June, October) annually from 2008 to 2012 to preserve the monocultures. The plots did not fertilized during cultivation. All the plots were harvested once a year in October.

2.2 Aboveground plant and belowground biomass sampling

Ten quadrats (1 m × 1 m) were randomly arranged in each plot in late August every year (2008-2012).

Aboveground biomass was measured by harvesting the upper plant parts, by (-clipping their stems at the soil surface), from ten each quadrats (1 m × 1 m) in each plot randomly in late August every year (2008-2012).

All green aboveground plant parts in each quadrat were collected separately by each individual species, and all the litters layer also were collected with the labeled envelops. Then these samples were dried at 105 °C until their mass was constant, and then their mass was weighed and recorded.

Belowground biomasses and soil samples were taken in the four corners and the center of the each

quadrats where were the aboveground biomass sampling points is (Fig. 1). Belowground biomass were was

collected using a soil drilling sampler with 9 cm inner diameter from at 0-100 cm soil layer (, and separated into increments every 10 cm). The roots in the soil samples were obtained by a 2 mm sieve. Then the remaining roots in the soil samples were isolated by shallow trays, and allowing the flowing water from the trays to pass through a 0.5 mm mesh sieve. All the roots samples were oven-dried at 65 °C then weighed.

2.3 Soil sampling and determination

The method of soil sampling is similar to belowground (using a soil drilling sampler with 4 cm inner diameter from 0-100 cm soil layer). In each quadrat, the same layer samples (every 10 cm) were mixed together and beto composed of a composite sample. The samples were passed through a 2-mm sieve to remove the roots and other debris. A 5 cm diameter and 5 cm high stainless steel cutting ring (~100 cm³) was used to measure soil bulk density (BD) at adjacent points to the soil sampling quadrats. Soil bulk density was measured at the depth of 0-100 cm (10 cm a layer then averaging). The dry mass were was measured after oven-drying at 105 °C. Soil organic carbon content was measured using the method of the vitriol acid-potassium dichromate oxidation (Walkley and Black, 1934). All the analyses of one sample were carried out in three replications.

2.4 Relative calculation

- BD was calculated depending on the oven dried weight of the composite soil samples (Deng et al., 2013).
- The SOC stock for each soil layer was calculated using the equation as follows (Deng et al., 2013):

$$C_s = BD \times SOC \times D/10 \tag{1}$$

- where, C_s is the SOC stock (Mg ha⁻¹); BD is the soil bulk density (g cm⁻³); SOC is the soil organic carbon content (g kg⁻¹); and D is the thickness of the sampled soil layer (cm).
- The SOC sequestration rate (SSR, Mg ha⁻¹ yr⁻¹) was calculated as follows (Hua et al., 2014):

144 SSR =
$$(C_t - C_0)/t$$
 (2)

where, $(C_t - C_0)$ is SOC sequestration; C_t is the SOC stock in 2012; C_0 is the SOC stock in 2008; t was

the duration of experiment (year).

The SOC sequestration efficiency was estimated using the SOC sequestration in the weight of total biomass (aboveground biomass and belowground biomass) of per unit area:

$$C_{se} = \triangle C / B_T / 10 \tag{3}$$

where, C_{se} is the SOC sequestration efficiency; $\triangle C$ (Mg ha⁻¹) is the SOC sequestration from 2008 to 2012; B_T (kg m⁻²) is the total biomass (above ground and below ground) from 2008 to 2012.

2.5 Statistical analyses

The data were examined for normality by the Shapiro-Wilk test and homogeneity of variances by the Levene test before analysis (Table 2). To get a normal distribution, performing statistical tests not normally distributed data were log-transformed. All data were expressed as mean values \pm standard error (M \pm SE). The means of SOC sequestration rate and SOC sequestration efficiency among the different grassland types were assessed using One-way Analysis of Variance (ANOVA). Two-way ANOVA of Type III was performed to test the influences of grassland types and time on SOC content, storage and bulk density. Tukey test was conducted to test the significance at p < 0.05 level. All the statistical analysis was performed with SPSS version 18.0 (SPSS Inc., Chicago, IL, USA).

3 Results

3.1 Aboveground net primary productivity

Between 2008 and 2012, the five cultivated grasslands in general had greater total biomass values than the uncultivated grassland and natural grassland (mean by 189.36%). In addition, the three grasslands cultivated with the leguminous species had greater annual total biomass than the two gramineous grasslands (mean by 72.6%), which lead to a greater total biomass values of the three leguminous species at the end of the study period. In particular, the L-MS grassland consistently had the greatest total biomass throughout the study period (Fig. 1a2a).

169 3.2 Soil SOC content and controls Results from two-way ANOVA showed that the plots types, year and interactions all significantly affected 170 171 total biomass, SOC content, and BD (p < 0.001, Table 53). The average SOC content followed leguminous grasslands $(4.21\pm0.31~{\rm g~kg^{-1}})$ > natural grasslands $(2.90\pm0.14~{\rm g~kg^{-1}})$ > uncultivated grassland (2.58 ± 0.17) 172 $g kg^{-1}$) > gramineous grassland $(2.46 \pm 0.15 g kg^{-1})$, and it increased over time in all grasslands (Table 24). 173 The L-MS grassland had the highest SOC content among the grasslands during the study period. The effects 174 of grassland types on soil bulk density followed uncultivated and natural grassland $(1.44 \pm 0.02 \text{ g cm}^{-3})$ > 175 gramineous grasslands $(1.43\pm0.01 \text{ g cm}^{-3}) > \text{leguminous grasslands} (1.40\pm0.01 \text{ g cm}^{-3}, \text{(Table 35)})$. 176 3.3 Soil organic carbon stock change 177 The SOC storage under all the grasslands increased significantly throughout the study period (Table 46), 178 179 with the three cultivated leguminous grasslands further significantly greater than those under the twogramineous grasslands. To be specific, in In the 0-20 cm soil layer, the SOC storage under the L-MS, L-CV 180 and L-OV grasslands increased from 9.73, 5.20, 7.27 Mg C ha⁻¹ to 14.95, 13.54, 12.05 by 5.22, 8.34 and 4.78 181 Mg C ha⁻¹, respectively, during the experimental period. 182 3.4 Soil carbon sequestration rate and sequestration efficiency 183 SOC sequestrations in three leguminous grasslands were greater than two gramineous grasslands (mean-184 by196.74%; Fig.1e). Three leguminous grasslands accumulated C with an average rate 1.00 Mg C-ha⁻¹-yr⁻¹ 185 which is more than the 0.34 Mg C - ha⁻¹- yr⁻¹ in gramineous grassland, and more than the average of 186 uncultivated and natural grasslands (0.25 Mg C - ha⁻¹- yr⁻¹; Fig.2c). 187 -The mean SOC sequestration efficiency in the leguminous grassland was about 0.26, which was 188 significantly greater than others grassland types (0.13; p<0.05; Fig. 1d2d). The maximum and minimum 189 190 efficiency values were 0.37, 0.08 in L-CV, G-PA grassland, respectively. The average SOC sequestrationefficiency in leguminous grassland was two times greater than gramineous grassland. 191

4 Discussion

Grassland has a significant effect in arid and semi-arid areas carbon cycle through changing soil carbon
accumulation rates and turnover, soil erosion, and vegetation biomass (Deng et al.,2014a; Liu et al., 2016a).
Plant regulated SOC stock by controlling carbon assimilation, transfer and storage in the plant root system,
then through plant respiration and leaching release from soil to atmosphere(De Deyn et al., 2008;
Garcia-Diaz et al., 2016). SOC inputs mostly originate from decaying aboveground and belowground plant
tissue, so greater soil C accumulation was mainly ascribed to increasing soil C input from higher biomass
production (Deng et al., 2014c; Wu et al., 2016). Mutualistic symbionts (N-fixing bacteria and mycorrhizal
fungi) are also an important source of carbon input to soil, especially in actively growing plants (Bardgett et
al., 2005). In grassland, atmosphere carbon was is sequestrated through photosynthesis and respiration, then
carbon fixing fixes in stable SOC pool or releasing release back into the atmosphere (Post et al., 2000).
Therefore, studying the carbon sequestration in grassland ecosystems can help to identify the magnitude of
global carbon sinks and sources (Li et al., 2014). Our results showed that leguminous grassland had greater
SOC content and storage efficiency than gramineous grassland. SOC content of all grassland plots showed
some differences between each other (Table 2 and 3). The average SOC content in leguminous grasslands-
was 2.64 g kg ⁻¹ and that in gramineous grasslands was 1.97 g kg ⁻¹ . Moreover, both soil bulk density of
leguminous and gramineous grasslands were 1.46 g cm ⁻³ in 2008. The reasons for the SOC content difference
result from precedent soil conditions and cultivated grasses. The Ddifferencet types of cultivated grasses, as
well as the precedent soil conditions are is probably the majortwo reasons for the SOC content differences
between leguminous and gramineous grasslands.
Different species may incorporate more or less carbon according to their specific metabolism. Legumes
have been identified as a key driver of C sequestration in many studies (Fornara and Tilman, 2008; Wu et al.,
2016). The irregular distribution of precedent plant residues and roots resulted in the patch of nutrients in the

soil and changing the soil physical conditions, such as SOC and BD. In addition, mutualistic symbionts-(N-fixing bacteria and mycorrhizal fungi) are also an important source of carbon input to soil, especially inactively growing plants (Bardgett et al., 2005). Symbiosis can increase plant productivity through enhancedthe acquisition of limited resources. Legumes live in a symbiosis with *Rhizobium* bacteria, which fix atmospheric N. Moreover, many previous studies had demonstrated that soil carbon and total nitrogen are significantly and positively correlated (Deng et al., 2013; De Oliveira et al., 2015). Therefore, mycorrhizal fungi can immobilize carbon in their mycelium and improve carbon sequestration in soil aggregates (Rillig and Mummey, 2006). It might be expected that the cultivated leguminous grasslands significantly improved soil N content that led to the greater carbon sequestration ability than the non-leguminous grasslands. Moreover, mycorrhizal fungi can immobilize carbon in their mycelium and improve carbon sequestration insoil aggregates (Rillig and Mummey, 2006). Our results demonstrated that a key variable associated withhigher SOC content in leguminous grasslands than gramineous grasslands is the greater total biomassaccumulation. The leguminous grasslands had both higher above- and belowground biomasses thangramineous grasslands. Total biomass was 16.35 kg m² in leguminous grasslands, which is 9.47 kg m² more than gramineous grasslands from 2008 to 2012. In addition, the grasslands in our study without grazing and only harvesting the aboveground biomass annually, so all the aboveground stubble and plant litters be inputto soil as a carbon supply. SOC mostly originates from decaying this aboveground and belowground planttissue, so greater soil C accumulation was mainly ascribed to increasing soil C input from higher biomassproduction (Deng et al., 2014c; Wu et al., 2016). Previous studies had showed that plant regulated SOC stock by controlling carbon assimilation, its transfer and storage in plant root system, then through plant respiration and leaching its release from soil to atmosphere (De Deyn et al., 2008). In addition, the biomass fraction resulting in SOC build-up (plant residuals) was strongly affected by management practices including the selection of plant species (Don et al., 2011). Species composition had a great role in determining the

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aboveground productivity (Liu et al., 2016b). Over relatively long time, the proportion of the aboveground
biomass enters soil as organic matter and incorporates into soil through physical and biological processes.
Some leachates from plant material in the litter layer, root exudates, solid decomposed litter and fragmented
plant structure materials were the main sources of soil organic matter (Jones and Donnelly, 2004; Novara et
al., 2015). The amount of plant residuals returned to the soil directly affected the SOC (Musinguzi et al.,
2015; Wasak et al., 2015), and mostly perennial plants were managed with high planting densities to produce
greater biomass exports (Hobbie et al., 2007; K öchy et al., 2015). Deng et al. (2014c) have found that plant
biomass is the key driver in soil carbon sequestration. In this study, the SOC increased dramatically in
leguminous grassland due to the greater total biomasses of the leguminous grasses, which resulting in and the
increased increasing soil carbon inputs from the litter layer and root biomass (De Deyn et al., 2008; Wu et al.,
2010; Novara et al., 2015). Moreover, Symbiosis can increase plant productivity through enhancing the
acquisition of limited resources. Our results demonstrated that a key variable associated with higher SOC
content in leguminous grasslands than gramineous grasslands is the greater total biomass accumulation. The
leguminous grasslands had both higher above- and belowground biomasses than gramineous grasslands.
Total biomass was 16.35 kg m ⁻² in leguminous grasslands, which are 9.47 kg m ⁻² more than gramineous
grasslands from 2008 to 2012. In addition, the grasslands in our study without grazing and only harvested the
aboveground biomass annually, so all the aboveground stubble and plant litters input to soil as a carbon
supply.
SOC sequestration rates in the cultivated leguminous grasslands were significantly higher than that in the
gramineous grasslands (Fig. 1c). This maybe resulted from SOC sequestration and the different
decomposition rates in soils, because the cultivated leguminous and gramineous grass species result in
multifarious nutrient conditions. The slower rates of decomposition might make soil carbon storages
increased faster in more nutrient poor soils (Vesterda et al., 2002; Deng et al., 2014a), L.CV grassland has

the highest SOC sequestration rate and efficiency but with the lowest total biomass among the leguminous grasslands. The reasons maybe the different species with the various C sequestrate capability, but the potential mechanism under each species need further studies to demonstrate. Leguminous grasslandsachieved greater SOC sequestration rates due to the total biomass was higher than that in the gramineousgrasslands. Litter and fragmented plant parts at the soil surface are decomposed by micro-organisms and are gradually incorporated into the soil through some complex processes, such as humification and mineralization (Novara et al., 2015). Legumes had the ability to develop root nodules and to fix nitrogen in symbiosis with compatible rhizobia, which should improve the soil nutrient status and microbial community. In addition to this, rates of decomposition in leguminous grassland may be higher due to excellent soil physical and chemical conditions. Root nodules promote the symbiosis with micro-organisms, which are responsible for the decomposition of the plants, and therefore constitute the key of the transmission of the stored carbon into the soil. Moreover, many previous studies had demonstrated that soil carbon and totalnitrogen are significantly and positively correlated (Deng et al., 2013; De Oliveira et al., 2015). Therefore, it might be expected that the cultivated leguminous grasslands had significantly improved soil N contents that led to a greater carbon sequestration ability than the non-leguminous grasslands. Furthermore, the resulting increase increasing in fertility of the soils in under the leguminous grasses should facilitate the increased increasing productivity of the plants. Our results showed that SOC sequestration efficiency under leguminous grasslands was evidently greater than that in the gramineous grasslands (Fig. 1d). It is noteworthy that L-MS grassland had the highest total biomass 22.59 kg m⁻² which is 2.38 times as much as the average of gramineous grasslands (Fig. 1a2a),). mMoreover, SOC sequestration in L-MS grassland is 3 times as much as the average of gramineous grasslands (Fig. 1+b2b). The differences of species biological characteristics determine the capacity of carbon sequestration. So the SOC sequestration efficiency in L-MS grassland is higher than gramineous grasslands.

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Despite the indications from this study of higher SOC sequestration rate and efficiency in leguminous grassland, specific researches are still needed to determine the potential mechanisms of each species in sequestrating carbon. Many studies have been demonstrated legumes are higher water consuming plants than gramineous in arid and semi-arid areas, so it is necessary to balance the ecological effect of grassland for rational utilization of resources.

Conclusion

Leguminous grasslands had greater SOC storage, sequestration rate and efficiency than gramineous grasslands. The greater soil C accumulation of leguminous grasslands was mainly ascribed to the capacity to incorporate carbon and the higher biomass production. Leguminous grasslands accumulated an average rate of 0.64 Mg C·ha⁻¹·yr⁻¹ more than gramineous grasslands. The average SOC sequestration efficiency in leguminous grasslands was 2 times greater than that in the gramineous grasslands. The Our results indicate that cultivated leguminous grasslands sequestered more soil carbon with a higher SOC sequestration efficiency than cultivated gramineous grasslands in arid and semi-arid areas. Our results provide a reference for ecological management in arid and semi-arid areas.

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Table 1. Description of studied grassland types.

Grassland types	Species	Seeding rates (kg-ha ⁻¹)			
Leguminous grassland	Coronilla varia L.	7.5			
	Onobrychis viciaefolia Scop.	30			
	Medicago sativa L.	12			
Gramineous grassland	Poa annua L.	7.5			
	Agropyron cristatum (L.) Gaertn.	15			
Uncultivated grassland	Abandoned cropland. Natural successional species were present, e.g., <i>Chenopodium album L.</i> , <i>Agropyron cristatum</i> L.				
Natural grassland	A local native grassland community. Dominant species were <i>Stipa breviflora</i> Griseb., <i>Stipa aliena</i> Keng, <i>Artemisia capillaris</i> Thunb., <i>Artemisia annua</i> L.				

Table 3 The p values of homogeneity of variances by the Levene test and normality by the Shapiro-Wilk test

in soil organic carbon content (SOC), soil C storage, and soil bulk density (BD).

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G 1 1		Levene tes	<u>t</u>	<u>S1</u>	hapiro-Wilk te	<u>st</u>
Grassland types	<u>BD</u>	SOC	SOC storage	BD	SOC	SOC storage
<u>L-MS</u>	<u>0.05</u>	<u>0.27</u>	<u>0.54</u>	<u>0.07</u>	<u>0.12</u>	<u>0.15</u>
<u>L-OV</u>	0.18	0.03	<u>0.04</u>	<u>0.19</u>	<u>0.32</u>	<u>0.36</u>
<u>L-CV</u>	0.17	<u>0.84</u>	<u>0.10</u>	<u>0.53</u>	<u>0.18</u>	<u>0.18</u>
<u>G-PA</u>	0.12	<u>0.10</u>	<u>0.01</u>	<u>0.07</u>	<u>0.03</u>	<u>0.02</u>
<u>G-AC</u>	0.02	0.09	<u>0.26</u>	<u>0.09</u>	<u>0.17</u>	<u>0.22</u>
<u>Un-G</u>	<u>0.01</u>	<u>0.10</u>	<u>0.06</u>	<u>0.03</u>	<u>0.05</u>	<u>0.03</u>
<u>Na-G</u>	<u>0.78</u>	<u>0.27</u>	<u>0.37</u>	<u>0.03</u>	<u>0.31</u>	<u>0.44</u>

Table 3 Two-way ANOVA F and p values for the effects of plot types, year, and interactions on total biomass

(TB), soil organic carbon content (SOC), soil C storage, and soil bulk density (BD). Bold numbers indicate

statistical significance.

<u>Factor</u>	<u>df</u>	<u>TB</u>		SOC		<u>C storage</u>		<u>BD</u>	
		<u>F</u>	<u>p</u>	<u>F</u>	<u>p</u>	<u>F</u>	<u>p</u>	<u>F</u>	<u>p</u>
Grassland types	<u>6</u>	<u>296.19</u>	<0.001	<u>40.52</u>	<u><0.001</u>	<u>42.03</u>	<0.001	<u>42.48</u>	<u><0.001</u>
<u>Year</u>	<u>4</u>	100.67	<u><0.001</u>	49.37	<0.001	41.05	< <u>0.001</u>	7.24	< <u>0.001</u>
<u>interaction</u>	<u>24</u>	<u>32.57</u>	< <u>0.001</u>	<u>2.30</u>	0.001	<u>2.00</u>	0.001	<u>7.36</u>	< <u>0.001</u>

Table 24. Soil C concentration (M \pm SE g kg⁻¹, average value of 100 cm soil depth) in different years and grassland types. Note: The grassland types were: L-Cv, *Coronilla varia*; L-Ov, *Onobrychis viciaefolia*; L-Ms, *Medicago sativa*; G-Pa, *Poa annua*; G-Ac, *Agropyron cristatum*; Un-G, uncultivated grassland; Na-G, natural grassland. Values followed by different lower-case letters within columns and upper-case letters within rows are significantly different at p<0.05.

Grassland types	2008	2009	2010	2011	2012
L-CV	2.31 ±0.04dE	3.09±0.05cD	4.22±0.04bC	4.91±0.02bB	5.92±0.05bA
L-OV	2.70±0.04bE	3.33±0.02bD	3.96±0.02cC	4.69±0.08cB	5.44±0.12cA
L-MS	2.92±0.06aE	3.62±0.05aD	4.38±0.02aC	5.55±0.09aB	6.13±0.05aA
G-AC	1.90±0.01gE	2.13±0.03fD	2.56±0.04eC	2.94±0.03eB	3.46±0.06dA
G-PA	2.03 ±0.01fE	2.14±0.02fD	2.26±0.02fC	2.57 ±0.01fB	2.65 ±0.02fA
Un-G	2.20±0.08eCD	2.35±0.02eC	2.42±0.04efC	2.81 ±0.01eB	3.16±0.02eA
Na-G	2.53±0.08cB	2.71±0.10dB	2.80±0.12dB	3.18±0.13dA	3.26±0.06eA

Table 35. Soil bulk density (M \pm SE g cm⁻³, average value of 100 cm soil depth) in different years and grassland types. Note: The grassland types were: L-Cv, *Coronilla varia*; L-Ov, *Onobrychis viciaefolia*; L-Ms, *Medicago sativa*; G-Pa, *Poa annua*; G-Ac, *Agropyron cristatum*; Un-G, uncultivated grassland; Na-G, natural grassland. Values followed by different lower-case letters within columns and upper-case letters within rows are significantly different at p<0.05.

Grassland types	2008	2009	2010	2011	2012
L-CV	1.41±0.01dAB	1.42±0.01bA	1.39±0.01cB	1.37±0.01cdC	1.35±0.01dD
L-OV	1.51±0.01aA	1.46±0.01aB	1.40±0.01bcdC	1.36±0.01dD	1.33±0.01eE
L-MS	1.47±0.15bcA	1.47±0.01aA	1.43±0.02abAB	1.39±0.02bcBC	1.36±0.01cC
G-AC	1.45±0.01cA	1.46±0.02aA	1.46±0.01aA	1.45±0.01aA	1.39±0.01bB
G-PA	1.47±0.01bcA	1.46±0.01aA	1.39±0.01cB	1.38±0.01cdC	1.34±0.01eD
Un-G	1.48±0.01bA	1.47±0.01aB	1.43±0.01abC	1.42±0.01bD	1.40±0.01aE
Na-G	1.49±0.01abA	1.48±0.01aB	1.42±0.01bcC	1.42±0.01bCD	1.41±0.01aD

Table 46. SOC stock (M \pm SE Mg $\underline{\mathbb{C}}$ ha⁻¹) at the depth of 0-100 cm in different years and grassland types. Note: The grassland types were: L-Cv, *Coronilla varia*; L-Ov, *Onobrychis viciaefolia*; L-Ms, *Medicago sativa*; G-Pa, *Poa annua*; G-Ac, *Agropyron cristatum*; Un-G, uncultivated grassland; Na-G, natural grassland. Values followed by different lower-case letters within columns and upper-case letters within rows are significantly different at p < 0.05.

Grassland types	2008	2009	2010	2011	2012
L-CV	31.49±0.31dE	43.10±0.60cD	57.92±0.87abC	66.67±0.17bB	79.34±0.80bA
L-OV	40.05±0.36bB	48.57±0.41bAB	55.41±0.41bAB	63.89±1.09bAB	72.66±1.38cA
L-MS	43.75±0.87aE	53.69±0.89aD	63.20±1.28aC	77.50±1.62aB	83.77±0.76aA
G-AC	27.11±0.27fE	30.87±0.60fD	37.10±0.60cC	42.53±0.33cB	48.10±0.82dA
G-PA	29.29±0.06eC	30.80±0.36fB	31.35±0.19dB	35.38±0.06eA	35.36±0.37fA
Un-G	32.03±0.65dD	33.83±0.18eC	33.83±0.52cC	38.72±0.17dB	43.25±0.22eA
Na-G	36.25±0.61cB	38.40±1.25dB	39.26±1.61cB	44.74±2.00cA	45.20±0.98eA

_Table 5 Two way ANOVA F and p values for the effects of plot types, year, and interactions on total biomass (TB), soil organic carbon content (SOC), soil C storage, and soil bulk density (BD). Bold numbers indicate statistical significance.

Figure captions

Fig.1 The studied site localization and schematic figure of the sampling strategies.

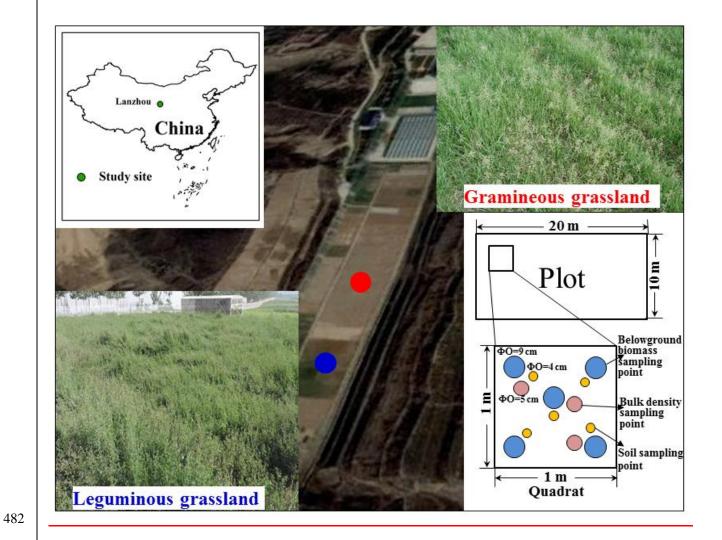


Fig. 21 Total biomass (a), SOC sequestration (b), SOC sequestration rate (c) and SOC sequestration efficiency (d) for different grassland from 2008 to 2012,. Note: The grassland types were: L-Cv, *Coronilla varia*; L-Ov, *Onobrychis viciaefolia*; L-Ms, *Medicago sativa*; G-Pa, *Poa annua*; G-Ac, *Agropyron cristatum*; Un-G, uncultivated grassland; Na-G, natural grassland. Bars indicate mean \pm standard error. Bars with the different lowercase letter above them indicate there was significant difference between the means at p<0.05 level. The dotted lines indicate the means of the same grassland types.

