



1 **The hidden ecological resource of andic soils in mountain ecosystems:**
2 **evidences from Italy**

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12

13 **Abstract**

14 Andic soils have unique morphological, physical and chemical properties that induce
15 both considerable soil fertility and great vulnerability to land degradation. Moreover
16 they are the most striking mineral soils in terms of large organic C storage and long C
17 residence time; this is especially related to the presence of poorly crystalline clay
18 minerals and metal-humus complexes. Recognition of these soils is then very important.
19 Here we attempt to show, through the combined analysis of 35 sampling points chosen,
20 throughout the Italian non volcanic mountain landscapes, in accordance to specific
21 physical and vegetation rules, that soils rich in poorly crystalline clay minerals have an
22 utmost ecological importance.



1 More specifically, in various non-volcanic mountain ecosystems (>700 m) and in low
2 slope gradient locations (<12°), in agreement to recent findings, we found the
3 widespread occurrence of soils with andic features having distinctive physical and
4 hydrological properties including low bulk density and remarkable high water retention.
5 Furthermore, we show a demonstration of the ability of these soils to affect ecosystem
6 functions by analysing their influence on the timescale acceleration of photosynthesis
7 estimated by NDVI measurements.

8 Our results are hoped to be a starting point for better understanding the ecological
9 importance of andic soils and also possibly to better consider pedological information in
10 C balance calculations.

11

12 **Keywords: fertile soils, high carbon storage capability, NDVI measurements,**
13 **hydrological properties, Andosols**

14



1 **1. Introduction**

2 Soils having andic features (allophanic and non-allophanic) are known to have a unique
3 set of soil morphological, physical and chemical properties. Between them (i) high
4 porosity (bulk density generally $< 0.90 \text{ g cm}^{-3}$), (ii) friable structure, (iii) high water
5 retention capacity, (iv) large reserves of easily weatherable minerals, (v) high
6 susceptibility to liquefaction, etc. Moreover, between all mineral soils, those with andic
7 features have the largest C storage capacity and long C residence time (Post, 1983;
8 Batjes, 1996; Amundson, 2001), which can be ascribed to the presence of poorly
9 crystalline clay minerals (Basile-Doelsch et al., 2005) and fungal and arthropodal SOM
10 (Nierop et al., 2005), but also to the specific physical and chemical properties that make
11 these soils some of the world's most fertile (Leamy, 1984; Shoji et al., 1993; McDaniel
12 et al., 2005). Despite these characteristics associated to C storage, andic features are
13 simply not considered in global carbon balance estimates (e.g. IPCC, 2006; Luo et al.,
14 2015); in fact in these estimates - in the best of cases - the contribution of soils (Parton
15 et al., 1987) is limited to organic C and soil texture parameters ignoring both other
16 important chemical and physical properties and the occurrence of well-known analytical
17 artefact in using texture data on soils difficult to disperse such as those having andic or
18 oxic features (Bartoli et al., 1991).

19 This lack of acknowledgment of andic soils is becoming more important considering
20 that in recent years soils with andic features have been found, along with well
21 established volcanic landscapes (Shoji et al., 1993; Arnalds & Stahr, 2004; Lulli, 2007),
22 in many “non-volcanic” mountain ecosystems (NVME) throughout the world (e.g.
23 Baumler et al., 2005; Dümig et al., 2008; Iamarino & Terribile, 2008; Scarciglia et al.,
24 2008; Graham & O'Geen, 2009; Rasmussen et al., 2010; McDaniel & Hipple, 2010;



1 Vingiani et al., 2014; Estevez et al., 2016). Given that two or three times more C is
2 stored in soils (Dixon et al., 1994) than occurs in the atmosphere as CO₂ and that andic
3 soils have such important C storage abilities (Torn et al., 1997), the above lack of
4 acknowledge of andic soils in carbon balance estimates is indeed unfortunate.

5 Moreover, in view of their large C storage capability, the danger of degradation of andic
6 soil is indeed high because they are some of the most vulnerable soils in the world in
7 terms of soil erosion (Arnalds, 2001) and rapid flow landslides (Basile et al. 2003;
8 Terribile et al. 2007; Vingiani et al., 2015).

9

10 **1.1. Aim and rationale**

11 All the above shows the need for a much better understanding about the importance of
12 andic soils and their ecological role. In this context, the aim of this contribution is to
13 attempt an insight about the influence of andic soil in Italian NVME over (i) vegetation,
14 through remotely sensed vegetation indexes and (ii) soil hydrological properties of
15 utmost importance for plant growth.

16 To achieve the above, a combined approach has been undertaken evaluating both 35
17 soils having different degrees of andic features in NVME (Figure 1) and the NDVI
18 dynamics of their sites.

19 All sites were chosen in order to select mountain soils (> 700 m asl) in conservative
20 geomorphological settings (slope gradient < 12°) and in areas with high primary
21 productivity (estimated using time series max NDVI value) from different parts of Italy
22 (see methods and Iamarino and Terribile, 2008).

23 The background of this approach being that (i) the above environmental factors can
24 promote andosolization and (ii) most importantly, that the great fertility of soils with



1 andic features positively affects plant primary productivity in natural ecosystems. Hence
2 the use of remotely sensed vegetation indexes (i.e. NDVI, EVI, etc.) can be a valuable
3 tool to address this topic: NDVI (Rouse at al., 1973) is strongly related to
4 photosynthetic activity and has been widely used to estimate landscape patterns of
5 primary production (Wang at al., 2004; Fensholt et al., 2012) and even net primary
6 production (Tucker & Sellers, 1986). Moreover, time series of NDVI and the related
7 NDVI metrics have proved to be a powerful tool for addressing plant dynamics and
8 yield prediction in both agriculture and natural ecosystems at different scales (Reed et
9 al., 1994; Zhang et al., 2003; Bolton & Friedl, 2013).

10

11 **2. Materials**

12 **2.1. Study site**

13 This specific work refers to the whole Italian mountain territory (Figure 1). Italy
14 develops between the 35° and 47° North parallel and it is located in the middle of the
15 temperate zone of the Northern Hemisphere. It has an extremely articulated territory; 2
16 major mountain chains occupy more than 35% of the entire national surface: (i) the
17 Apennines, with predominantly sedimentary rocks, crossing almost entirely the Italian
18 territory from S to N, with altitude reaching 2900 m asl (Gran Sasso); (ii) the Alps,
19 having predominantly metamorphic and igneous rocks, separating Italy from the rest of
20 Europe, with maximum altitude over 4,000 m asl (Monte Bianco, Monte Rosa,
21 Matterhorn). The remaining territory is mainly occupied by hilly systems (about 40%)
22 including those portions of Apennines slowly degrading towards the sea, both at E and
23 W. Plain systems only occupy just over 20% of the entire territory.



1 In general terms the climate - known to be mild – is heavily influenced by the sea. With
2 respect to Italian mountain areas it can be assumed that for soil climate (Soil Survey
3 Staff, 2014) the mean moisture regime is udic (it may become ustic at lower elevation)
4 whereas the mean temperature regime is generally mesic (it may become frigid and
5 cryic at high elevation) (Costantini et al., 2004; Costantini et al., 2013).

6

7 **2.2 Soil sampling**

8 Soil sampling was designed to collect fertile mountain soils in conservative
9 geomorphological settings from different parts of Italy. The soils were sampled from (i)
10 mountain environments (> 600 m asl estimated by a 270 m spatial resolution DEM
11 obtained from the Italian Geological Service), (ii) geomorphological conservative
12 landscapes with moderately low slopes (slope gradient < 30° evaluated by the DEM) to
13 minimise the risk of sampling eroded soils and finally (iii) areas with high primary
14 productivity estimated using the max NDVI value (NDVI threshold 0.65) obtained from
15 MODIS Images MVC (230 m spatial resolution) for the period 28/7 - 13/8 2014 (which
16 is a strong vegetative growth period in Italy). Morphological and chemical (aggregated)
17 data of these pedons (28 soils after the selection reported in paragraph 2.3) along with
18 the background to this methodology are given in Iamarino and Terribile, 2008.

19 These information were further supplemented with data of 7 soils: 4 newly surveyed
20 and analysed soils, 3 soils reported in the scientific literature and consistent with the
21 previously stated rules: 1 soil concern research work in the Abruzzo region (Frezzotti
22 and Narcisi, 1996), 1 soil the CON.ECO.FOR program (Corpo Forestale dello Stato,
23 2003), a further soil was retrieved from the ISRIC database (ISRIC, 2005).

24



1 **2.3 NDVI and land use data**

2 In-depth analysis on time-based NDVI was performed using a MODIS VI algorithm
3 which operates on a per-pixel basis and relies on multiple observations over a 16-day
4 period to generate a maximum composite VI MVC. In order to extract the NDVI
5 metrics (maximum NDVI, integrated NDVI sum over the growing period, acceleration
6 of photosynthesis or rate of green-up, NDVI derivatives) some pre-processing of the
7 data were necessary (i.e. cloud contamination) following established procedures (Reed
8 et al., 1994). After such processing, about 15% of the NDVI observations had to be
9 discarded and the whole dataset was excluded from this work. This is related to well-
10 known problems in remote sensing, due to high and persistent cloud contamination and
11 in some cases also to the presence of rock outcrops inside the area of the investigated
12 pixels.

13 NDVI data were chosen to incorporate years having marked contrasting climate and
14 then - potentially - contrasting vegetation indexes trends and metric. Analysing the
15 climatic database published by the Italian Ministry of Environment for the whole
16 country (<http://www.isprambiente.gov.it/>), we have chosen years 2003, 2005, 2014.
17 These climatic years have the following trends (values below are ranked in the order
18 2003, 2005, 2014 respectively):

- 19 - similar yearly mean temperature: 13°C, 12°C, 13°C;
- 20 - evident differences in yearly mean maximum temperature, 36°C, 35°C, 33°C;
- 21 - most importantly, marked differences in yearly cumulated rainfall, respectively
22 766 mm (SD: 172mm), 870 mm (SD: 231 mm), 1143mm (SD: 540mm);
- 23 - marked differences in Standardized Precipitation Index (McKee et al.,1993),
24 varying in the range 0.5-0.5; 0.5-0.0; 1.0-2.0. This index is a well-known simplified



1 indicator for monitoring drought and periods of anomalously wet events and it shows
2 evident droughts for years 2003 and 2005.

3 The Corine land cover (CLC level 4, 5) classification (ISPRA, 2012) was used to
4 produce a preliminary evaluation of the main land covers. Corine land cover classes
5 were locally validated for each of the sampled sites. The reported land cover classes of
6 chestnut, beech and broadleaf oak must be considered classes of land cover where these
7 species are predominant but not exclusive. The grassland class refers to both continuous
8 and discontinuous natural grassland.

9

10 **3. Methods**

11 All statistical analysis was performed using two-tailed tests; ANOVA (Tamhane
12 method) was performed for multiple comparisons of means. The reported test of
13 significance for the latitude was performed on a “metres from the equator” basis.

14 At each site a soil profile was opened up, described (FAO, 1990) and sampled. Bulk
15 samples were collected from all the soil horizons for chemical analysis. Undisturbed
16 soil samples for hydrological analysis were collected from the main horizons with steel
17 cylinders of about 200 cm³.

18 Bulk samples after air drying were sieved to less than 2 mm and analysed (USDA-
19 NRCS, 2004): organic matter was determined by the Walkley & Black method;
20 Al/Fe/Si in the amorphous oxides/hydroxides and in the organic matter were extracted
21 respectively with ammonium oxalate (Fe_o, Al_o, Si_o) treatment at pH = 3 (Schwertmann
22 method) and their content levels were determined by ICP-AES. Values of Al and Fe
23 extracted with ammonium oxalate were used to calculate the andic feature index
24 $Al_o + 0.5Fe_o$. Phosphate retention was determined according to the method of Blakemore.



1 In order to simplify the comparison between soils features and land use or NDVI
2 metrics it was necessary to aggregate chemical data obtaining a single representative
3 value for the whole soil; then the contents of $Al_0+0.5 Fe_0$, P retention and organic
4 carbon were weighted according to horizon thickness for each of the pedons. Soils were
5 classified using the World Reference Base (IUSS Working Group WRB, 2015).

6 With respect to hydrological analysis, ten experimental points of the soil water retention
7 curve $\theta(h)$, ranging from saturation to -30 kPa of potential, were determined through use
8 of the tension table and 5 points at -100, -500, -800, -1200 and -1500 kPa were
9 determined through use of a pressure plate apparatus (Dane and Hopmans, 2002). The
10 soil samples were then dismantled and dried for 24 h in the oven at 105°C in order to
11 determine the water content from the weight data set and the bulk density.

12 The water retention experimental data were parameterised according to the unimodal
13 $\theta(h)$ relationship proposed by van Genuchten (1980), expressed here in terms of the
14 scaled water content, S_e , as Equation (1) below:

$$S_e = \left[1 + (\alpha|h|)^n \right]^{1/n-1} \quad (1)$$

15 with $S_e = (\theta - \theta_r) / (\theta_0 - \theta_r)$, and in which α (cm^{-1}) and n are curve shape parameters. θ_0 and θ_r
16 respectively represent the saturated water content (at $h=0$) and the residual water
17 content, and may either be fixed or treated as parameters to be optimized.

18 To obtain a synthetic description of water retention for an easy comparison with soil
19 chemical analysis, we used a numeric index (*IRI*) integrating the whole water retention
20 function (Basile et al., 2007).

21 The Integral Retention Index, *IRI*, is defined by:



$$IRI = \frac{1}{wp} \int_0^{wp} \theta d(\log_{10} |h|) \quad (2)$$

1 where $wp=4.2$ is the wilting point. This adimensional index ($0 < IRI < 1$) represents the
2 average value of the function $\theta(\log_{10} |h|)$ on the interval $[0, wp]$ and allows simple
3 comparisons of the whole water retention by coalescing it in a single characteristic
4 value.

5

6 **4. Results and Discussion**

7 **4.1 Soil and landscape**

8 The outcome results of our procedure in terms of soil analysis and WRB soil
9 classification (IUSS Working Group, 2015) show that Andosol and Cambisol alone
10 account for more than 80 % of the observations and, most interestingly, despite
11 differences in soil classification, in the vast majority of cases (about two-thirds) there is
12 a quite high content of poorly ordered clay minerals as estimated by $Al_o+0.5Fe_o$ % as
13 given in Figure 2 (moderate and well expressed andicity). Iamarino and Terribile (2008)
14 have reported further details (data reported as horizon-based means) on 42 of these
15 pedons proving the general absence of podsolization and depicting a scenario where
16 andosoliation is the main soil process.

17 In Table 1 are reported the main geographical and land cover features of the studied
18 soils along with NDVI metrics over three contrasting climatic years; the dataset shows
19 that Andosols, Cambisols and Phaeozems occur at similar latitudes and elevations and
20 beech, oak, chestnut and grassland are the main land use. More specifically, the main
21 land cover unit associated with Andosols and Cambisols is the beech forest but they also
22 occur in other land uses. Phaeozems are mostly associated with grassland.



1 In all years, in sites where Andosols occur the mean value of max NDVI, integrated
2 sum of NDVI and NDVI green-up is always the highest, as compared to other soil
3 classes. This finding is very interesting and it is consistent with the high fertility of
4 these soils. NDVI max and NDVI integrated sum (Jun-Aug) show significant
5 differences between the different land cover classes, following clear diversity in plant
6 biology.

7 The analysis of NDVI trend between the 3 investigated years, shows that, as expected,
8 NDVI max and NDVI sum values in the wetter 2014 are always higher than in the drier
9 2003 and 2005. Differently the NDVI green-up values are typically higher in 2003-2005
10 as compared to 2014 and this NDVI green-up difference is even more pronounced
11 moving towards the most andic soils (Andosols). All the above clearly suggest that soils
12 with andic features – typically having higher water storage as compared to other soils –
13 enabled to produce an higher green-up. Here we must also add that further analysis
14 would be required to evaluate at each site trends in soil water storage and temperature
15 before the green-up phase.

16 In Table 2 are reported the main features of the studied soils; the soil dataset shows that
17 all soils are deep, have a friable granular/crumb soil structure at the surface; moreover,
18 organic C, andic features (always $Al_0+0.5Fe_0 \geq 0.4\%$) and P retention range from
19 moderate to high. Of all the soils, Andosols, have the highest (i) soil depth, (ii)
20 $Al_0+0.5Fe_0$ % (weighted mean) and (iii) P retention % (weighted mean). Phaeozems
21 have the highest organic C (weighted mean) content.

22 Though $Al_0+0.5Fe_0$ and P retention values in Andosols differ significantly, there are no
23 such significant differences between the various land cover classes, suggesting that
24 vegetation is of little importance in determining andic features.



1 In general terms the investigated soils can be considered rather homogeneous in their
2 morphological, chemical and physical properties although they occur in very diverse
3 geological and climatic mountain ecosystems; a preliminary cautious estimate
4 (Iamarino, 2005) of their distribution in Italy has shown their presence on about 7×10^5
5 ha.

6 This finding parallel similar ones in other parts of the world where mountain soils with
7 andic features (not necessarily Andosols) have been reported in Bhutan (Baumler et al.
8 2005), in Brasil (Dumig et al., 2008), in California (Graham & O'Geen, 2010;
9 Rasmussen et al., 2010), Pacific North-West USA (McDaniel & Hipple, 2010), NW
10 Spain (Estevez et al., 2010) and also in Italy (Iamarino & Terribile, 2008; Scarciglia et
11 al., 2008; Vingiani et al., 2014).

12

13 **4.2 Andic features and soil hydrology**

14 Given the finding on the importance of andic soils (albeit not Andosols) in Italian non-
15 volcanic uplands, the question is raised as to whether such andic features are also
16 connected to those physical properties considered of key importance for plant growth,
17 namely bulk density and water retention due to their crucial role in water availability. In
18 order to address this issue, a selection of undisturbed soil samples, from horizons A and
19 B, of the previously investigated soils were analysed. The data (in Table 3) clearly show
20 the occurrence of very porous soils (low bulk density) and very high water retention
21 capability over the complete range of pressure head values. Surface A horizons
22 generally have lower bulk density and higher water retention explicit by IRI than the
23 subsoil B horizons, which must be ascribed to the contribution of organic carbon in



1 improving the soil structure (Kutilek and Nielsen, 1994) and therefore increasing water
2 retention and decreasing bulk density.
3 The positive high correlation (Figure 3) between $Al_o+0.5Fe_o$ (%) and IRI indicates that
4 higher andic features correspond to higher integrated water retention, hence very good
5 soil physical properties. This result is already established (Basile et al., 2007) but only
6 for soils having $Al_o+0.5Fe_o$ (%) larger than 2% while there are no positive evidence for
7 soils having much lower $Al_o+0.5Fe_o$ content (e.g. in the range 0.4-2.0%). All the above,
8 emphasises that poorly ordered clay minerals greatly affect soil physical properties even
9 at moderate to low concentration, which in turn could greatly affect water storage and
10 then water availability for plant ecosystem growth.
11 Such finding is important because it does not refer to soils in a unique location but
12 rather to a large variety of soils developed at different latitude and over different
13 bedrocks and land uses.

14

15 **4.3 Andic features and elevation against NDVI metrics**

16 To investigate this question further, bivariate correlation (Table 4) and regression
17 analyses (Figure 4) were performed between andic features ($Al_o+0.5Fe_o$ %) and NDVI
18 metrics for each of the observed land cover classes. In the vast majority of climatic
19 years and land cover classes, andic features have a positive correlation with NDVI
20 metrics but, generally, not significant for (i) NDVI max value and (ii) integrated sum of
21 NDVI (Table 4). By contrast, rather astonishing, andic features are always well
22 correlated with the rate of green-up (1st derivative of NDVI); this correlation is
23 significant for the driest years 2003 and 2005 and not for the wettest 2014. Highest
24 significant correlations are found when each land use is considered separately. For



1 instance, in 2003 the r Pearson between andic features and green-up is 0.82 for beech
2 and 0.83 for grassland while in the year 2005 is 0.86 for beech and 0.90 for grassland.
3 These results show that beech and grassland are the best performing to show the
4 ecological importance of andic features; furthermore, the data producing this high
5 correlation are spanned along a high range of $Al_0+0.5Fe_0$ % values (see Figure 4). This
6 performance could be explained considering that i) beech and grassland are more
7 spatially homogeneous land uses as compared to oak broadleaves (e.g. oak land use is
8 more heterogeneous being a potential mixture of very different species sometime even
9 including grassland); (ii) beech and grassland land uses are less affected by strong land
10 management practices as compared to chestnut (in fact in the Italian landscape it is often
11 managed as coppice); (iii) moreover it is well known that beech is very susceptible to
12 severe water stress (Teissier et al., 1981).

13 All the above can could well explain the more responsive NDVI signal of beech and
14 grassland to water stress as compared to oak broadleaves and chestnut.

15 To the authors best knowledge, it is the first time that it is shown a close connection
16 between NDVI metrics and soil andic features. This result can have important
17 consequences in terms of better understanding the ecology of Italian mountain
18 ecosystems.

19 Differently in many different environment often it has been reported the positive
20 variation of NDVI against elevation (Zhan et al. 2012; Walsh et al., 2001; Chen et al.,
21 2006), thus since soils with andic features occur in mountain areas, it was important to
22 test whether the observed relationship between NDVI metrics and andic features hinder
23 a possibly even closer relationship between NDVI metrics and elevation.



1 To this respect, table 4 shows that the correlation between NDVI metrics and elevation
2 is much more confusing with much lower r values as compared with those between
3 NDVI and andic features. Overall both the low and negative r values between many
4 NDVI metrics and elevation show that altitude (and possibly its covariates, i.e.
5 temperature and rainfall) do not adequately explain variations in green active biomass
6 parameters. Moreover, r values between andic features and elevation show very low
7 values (e.g. $r = 0.16$ for all sites) and do not show any consistent trend (data not shown).
8 Then here we can state that for the first time it has been demonstrated the ecological
9 importance of soils with andic features over different land use canopies with respect to a
10 large part of Italian mountains; most probably this finding has to be connected to the
11 unique hydro-pedological properties of these soils. In fact, this result is especially
12 evident in the driest years (2003, 2005) while is less important in the wettest 2014 year
13 thus it is rather evident that the water storage of these soils may play a key controlling
14 role.

15 Our findings are also important to better acknowledge the occurrence and the
16 importance of these soils in C sequestration/storage estimates. Indeed, deep andic soils
17 (as reported in this study) have about twice (Batjes, 1996) the mean organic C content
18 of deep Regosols, Cambisols and Podzols which previous soil inventories (Mancini,
19 1966; EuDASM, 2007) considered as the main soil types in the investigated landscapes.

20

21 **5. Conclusions**

22 Our study shows a close relationship between the degree of andic features and NDVI
23 metrics and especially with metrics describing acceleration of photosynthesis (green-



1 up). This finding demonstrates that there is yet much to be understood about the
2 ecological importance of soils in mountain ecosystem, at least for the Italian territory.
3 Moreover the acknowledge of the importance of these soils may also have important
4 consequences in terms of both soil protection in mountain environment (andic soil are
5 known to be easily erodible) and for better understanding the impact of climate change.
6 To this respect this study suggest that the unique water retention features of the andic
7 soils plays an important ecological role when comparing contrasting climatic years.
8 The above result are maybe even more pronounced considering that the current study
9 employed a rather simplified NDVI approach including data at coarse resolution
10 (MODIS) and no algorithm to mitigate the well-known saturation effect of NDVI
11 (Buschmann and Nagel,1993). Thus it is likely that in future, better focused studies,
12 may demonstrate even better and closer relationships between andic soils and green
13 biomass indicators.
14 Generally our results indicate the large potential in using remote sensed vegetation
15 index metrics to ameliorate soil spatial inventories. A question still arises as to whether
16 the general absence of strong significant correlation between andic features with both
17 “NDVI max” and “integrated NDVI sum” may be caused by the quoted NDVI
18 saturation effect.
19 Considering our results, it is also important to emphasise that the importance of andic
20 features in affecting ecosystem function is undoubtedly poorly expressed by soil
21 classification: in fact strict classification rules dealing with how/where to expect “andic
22 properties” (WRB: starting within 25 cm from the soil surface; Soil Taxonomy: within
23 60 cm) can lead to non-Andosols with very high andic features. However, andic
24 features, rather than soil class criteria, seem to better explain variability in NDVI



1 metrics and plant ecosystem dynamics and this finding must be of major concern for
2 ameliorating soil classification.

3 Although the importance of this key mineral soil in Italian mountain ecosystems is
4 demonstrated producing in turn large organic C storage and long C residence time,
5 proper implementation of these new data in terms of C balance calculation, reducing
6 uncertainties in carbon sequestration estimates and carbon sink national ecosystems
7 inventory, is indeed a major issue to be addressed.

8 Moreover, the given wide recognition of andic soils has important consequences both in
9 terms of C sequestration potentialities and C lost risks associated to this finding.

10 Suitable land management techniques are then required to match the exclusive
11 properties and problems connected to the presence of these soils.

12 Considering the many recent finds of “andic” soils worldwide, it is of great importance
13 to ascertain whether a wider occurrence of this hidden resource apply also to mountain
14 environments in other parts of the world.

15 Finally, we must emphasise that this study – focused on only 35 points over the Italian
16 landscape – is the methodological basis for producing statement at the national scale
17 where, accordingly, much more data are indeed required.

18

19



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FIGURE CAPTIONS

Figure 1. Location of the sampling points (black triangles).

Figure 2. Soil type (WRB classification) plotted against andic features estimated by $Al_o+0.5Fe_o$ % (weighted mean according to horizons thickness for each of the studied pedons).

The value of 0.4% in $Al_o+0.5 Fe_o$ is the “key out” requirement for entering in the Andosol (and/or Andisol) classes both in WRB and Soil Taxonomy classifications. The andic features estimated by $Al_o+0.5 Fe_o$ % can be considered weak in the range 0.4-1.0, moderate in the range 1.0-2.0 and well-expressed over 2.0.

Figure 3. Scatterplot between andic features estimated by $Al_o+0.5 Fe_o$ % and Integrated Retention Index (IRI). Coefficient of determination R^2 along with the number of data points (n) are reported.

Figure 4. Scatterplot between andic features estimated by $Al_o+0.5 Fe_o$ % (weighted mean $Al_o+0.5Fe_o$ % according to horizons thickness for each of the studied pedons) and the maximum value of the NDVI derivative. From left to right: grassland, beech, oak and chestnut. From bottom to top: year 2003, 2005 and 2014. The dashed lines show the linear regression for each land cover. Coefficient of determination R^2 along with the number of data points (n) are reported for each panel.



Figure. 1

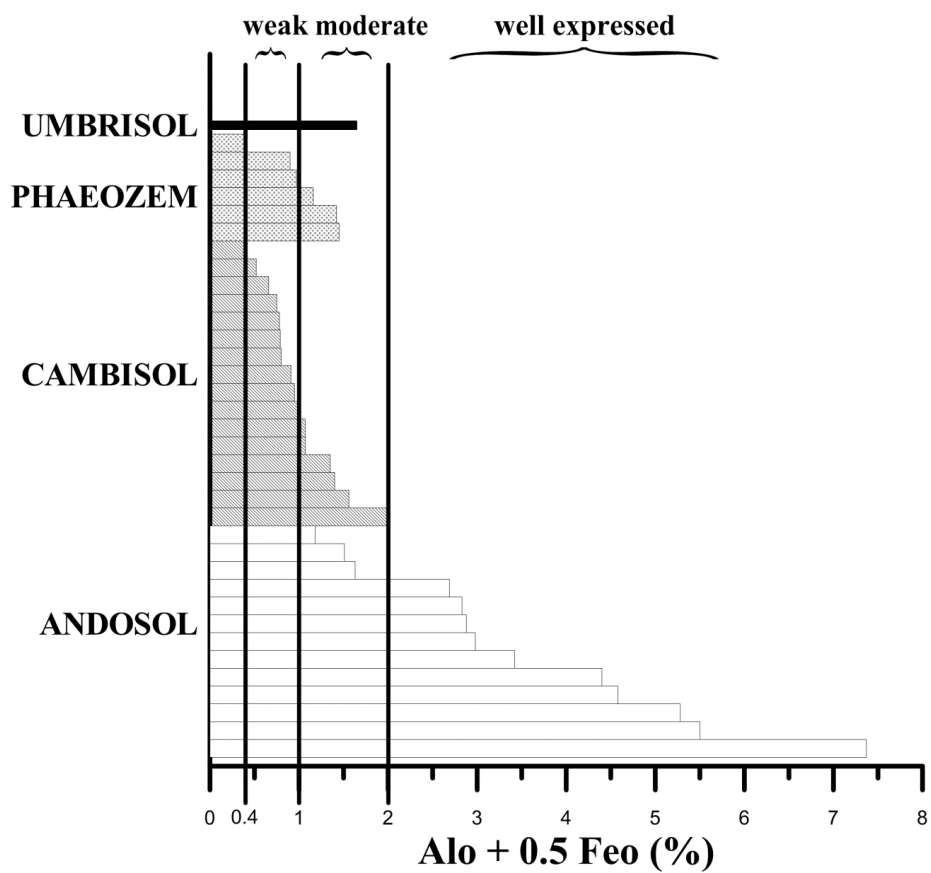


Figure 2

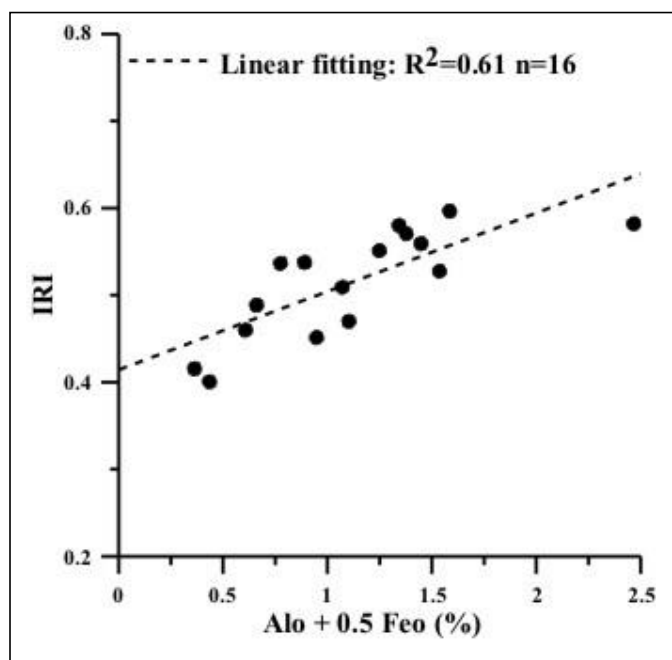


Figure 3

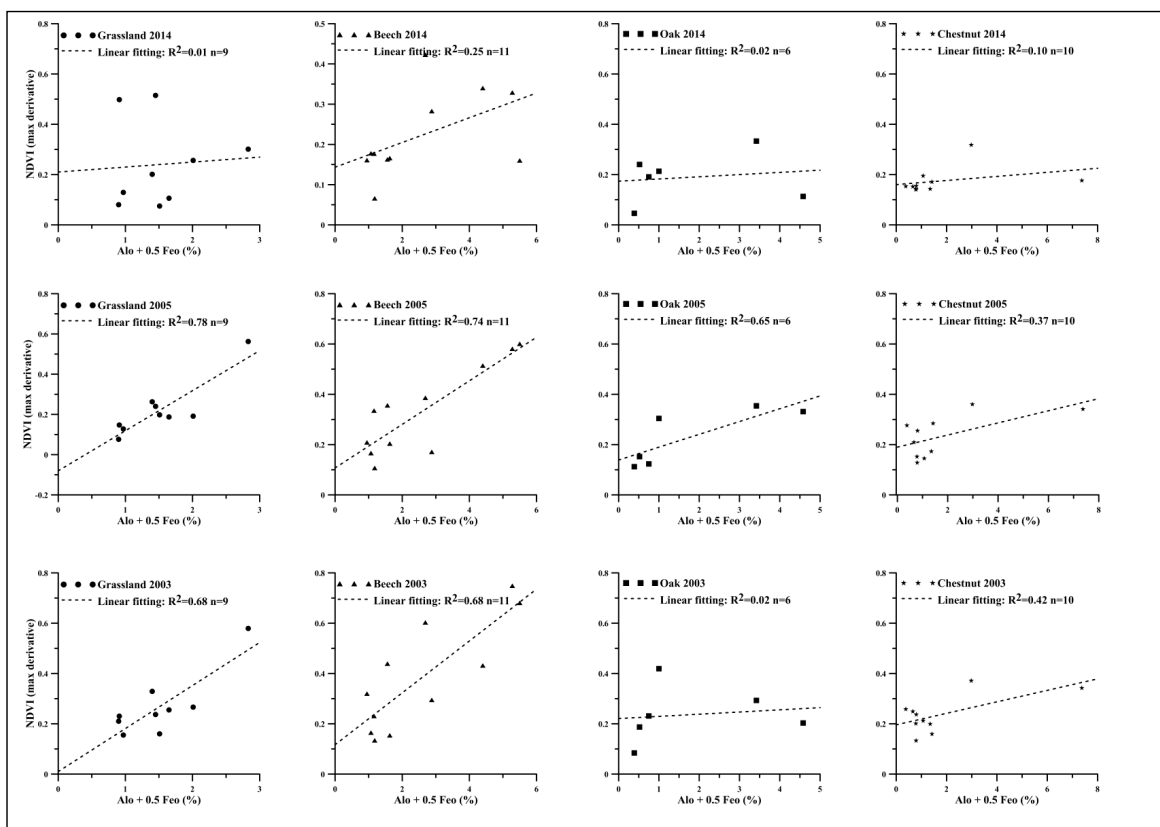


Figure 4



Table 1. Main geographical and land cover features of the studied soils. Abbrev.

Soils/ Land Cover	n.	Latitude mean	Elevation mean	Land Cover/ Soils mode	NDVI max mean		NDVI sum (Jun-Aug) mean		NDVI green-up (1 st derivative) mean				
					2003	2005	2003	2005	2003	2005	2003	2005	
All soils	35	4714218 N ± 187375	1006 ± 427	Beech	0.87 ±0.07	0.86 ± 0.09	0.88 ± 0.07	4.85 ±0.71	4.86 ± 0.77	5.06 ±0.49	0.29 ± 0.16	0.26 ± 0.14	0.21 ± 0.17
Andosols	13	4565657 N ± 144152	1040 ± 400	Beech	0.89 ±0.04	0.90 ±0.04	0.90 ± 0.05	5.07 ±0.41	5.11 ± 0.49	5.24 ±0.33	0.38 ±0.21	0.36 ±0.16	0.24 ± 0.12
Cambisols	16	4828328 N ± 103489	943 ± 474	Beech	0.87 ±0.07	0.86 ±0.09	0.88 ± 0.06	4.97 ±0.57	4.97 ± 0.67	5.08 ±0.50	0.25 ±0.08	0.20 ±0.07	0.20 ± 0.09
Phaeozems	6	4731804 N ± 239417	1100 ± 427	Grassland	0.80 ±0.09	0.80 ±0.12	0.82 ± 0.08	4.02 ±1.01	4.03 ± 1.06	4.62 ±0.55	0.18 ±0.06	0.20 ±0.10	0.19 ± 0.17
Beech	11	4630565 N ± 199235	1219** ±291	Cambisols Andosols	0.92 ±0.02	0.92 ±0.02	0.92 ±0.02	5.28** ±0.17	5.36** ± 0.13	5.41** ±0.10	0.38 ±0.22	0.33 ±0.18	0.22 ± 0.11
Chestnut	10	4829743 N ± 187599	680** ± 240	Cambisols Andosols	0.90 ±0.02	0.90 ±0.02	0.90 ±0.01	5.20** ±0.16	5.21** ± 0.11	5.23** ± 0.08	0.24 ±0.07	0.23 ±0.08	0.17 ±0.05
Oak broad.	6	4610094 N ± 139918	728 ± 424	Cambisols Andosols Phaeozems	0.86 ±0.07	0.86 ±0.06	0.88 ±0.05	4.77 ±0.71	4.91 ± 0.57	5.01 ±0.37	0.24 ±0.11	0.23 ±0.11	0.19 ±0.10
Grassland	8	4762927 N ± 111922	1330 ± 392	Cambisols Andosols Phaeozems	0.77 ±0.09	0.75 ±0.10	0.79 ±0.08	3.88** ±0.66	3.70** ± 0.74	4.40 ±0.56	0.27 ±0.14	0.23 ±0.15	0.26 ±0.17

** $\alpha < 0.01$, * $\alpha < 0.05$ (two-tailed test).

Abbr. n.: number of observations, broad.: broadleaf species.

The symbol ± after the mean value shows the standard deviation. The (n.) values refer to the number of observation available for NDVI analysis (see methods); in some sites because of strong cloud contamination not all the data could be used for NDVI analysis.

The upper part of the table refers to soil types (WRB) and the lower part refers to land cover (CORINE Land Cover classes) after site validation. NDVI MODIS metrics referring to a whole 2003, 2005, 2014 time series (16 day step).



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Table 2. Main soil features of the studied soils

Soils / Land Cover	n.	Soil depth	Structure of surface A horizon	Organic C	Alo+0.5Feo	P Retention	
		(solum)		mean	mean	mean	
		mean	mode	%	%	%	
		cm					
All soils	35	88 ± 37	Friable Gr. Cr. medium	37	38.0 ± 23.0	2.0 ± 1.7	62.9 ± 26.0
Andosols	13	115** ± 34	Friable Gr. Cr. medium	69	45.3 ± 26.6	3.6** ± 1.8	90.2** ± 14.6
Cambisols	16	75 ± 31	Friable Gr. Cr. fine; Cr. coarse	21	27.3 ± 15.1	1.0 ± 0.4	46.9 ± 17.2
Phaeozems	6	66** ± 21	Friable Gr. Cr. medium	57	50.9 ± 22.8	1.0 ± 0.4	49.1 ± 16.5
Beech	11	102 ± 28	Friable Gr. Cr. medium	41	40.6 ± 22.9	2.6 ± 1.7	83.6 ± 21.6
Castanea	10	95 ± 38	Friable Gr. Cr. Coarse	40	23.5 ± 10.8	1.8 ± 2.1	42.7 ± 21.9
Oak broad.	6	70 ± 55	Friable Gr. Cr. Fine	25	34.2 ± 22.3	1.8 ± 1.8	61.3 ± 27.5
Grassland	8	75 ± 25	Friable Gr. Cr. medium	50	55.6 ± 25.4	1.5 ± 0.7	57.5 ± 16.7

** $\alpha < 0.01$, * $\alpha < 0.05$ (two-tailed test).

Abbr. n.: number of observations, broad.: broadleaf species, Gr.: granular, Cr.: crumb, fine: < 2 mm, medium: 2-5 mm, coarse: 5-10 mm, very coarse: > 10 mm.

The symbol ± after the mean value shows the standard deviation.

The upper part of the table refers to soil types (WRB) and the lower part refers to Land Cover (CORINE Land Cover classes) after site validation. Chemical analyses are integrated over soil depth (solum).

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Table 3. Main physical parameters of selected soil horizons

Horizons	Mean Bulk Density		Mean WC at pF=4.2		Mean WC at pF=0		IRI	
	n.	g cm ⁻³	n.	cm ³ cm ⁻³	n.	cm ³ cm ⁻³		n.
All	35	0.87 ± 0.21	83	0.25 ± 0.09	16	0.79 ± 0.10	16	0.51 ± 0.06
A	16	0.79 ± 0.17	55	0.27 ± 0.09	7	0.85 ± 0.07	7	0.55 ± 0.04
B	19	0.93 ± 0.22	27	0.19 ± 0.07	10	0.75 ± 0.10	10	0.48 ± 0.06

Abbr. n.: number of observations, WC: volumetric water content, IRI: integrated water retention index.

The symbol ± after the mean value shows the standard deviation.

The table reports for soil horizons A and B mean bulk density, water retention at two different values of pF (0 and 4.2) corresponding to the pressure head of -0.1 and -1500 kPa, respectively, and the integrated retention index (IRI) which coalesces the water retention curve in a single value (Basile et al., 2006).



Table 4. Bivariate correlation analysis

	Alo+0.5 Feo (%)											
	2003				2005				2014			
	Mean NDVI max sum (Jun-Aug)	Mean NDVI green-up (1 st derivative)	Mean NDVI max	Mean NDVI sum (Jun-Aug)	Mean NDVI max	Mean NDVI green-up (1 st derivative)	Mean NDVI sum (Jun-Aug)	Mean NDVI max	Mean NDVI green-up (1 st derivative)	Mean NDVI sum (Jun-Aug)	Mean NDVI max	Mean NDVI green-up (1 st derivative)
All sites (n.:35)	0.19	0.19	0.17	0.20	0.17	0.61**	0.20	0.16	0.71**	0.20	0.20	0.23
Beech	0.36	-0.09	0.24	0.20	0.24	0.82**	0.20	0.42	0.86**	0.60*	0.50	0.50
Oak	0.16	0.28	0.12	0.29	0.12	0.14	0.29	0.20	0.81	0.28	0.15	0.15
Chestnut	-0.01	-0.21	-0.13	-0.004	-0.13	0.65*	-0.004	-0.21	0.61	-0.01	0.32	0.32
Grassland	-0.46	0.26	-0.01	0.02	-0.01	0.83*	0.02	-0.35	0.90**	-0.31	0.10	0.10

	Elevation											
	2003				2005				2014			
	Mean NDVI max sum (Jun-Aug)	Mean NDVI green-up (1 st derivative)	Mean NDVI max	Mean NDVI sum (Jun-Aug)	Mean NDVI max	Mean NDVI green-up (1 st derivative)	Mean NDVI sum (Jun-Aug)	Mean NDVI max	Mean NDVI green-up (1 st derivative)	Mean NDVI sum (Jun-Aug)	Mean NDVI max	Mean NDVI green-up (1 st derivative)
All sites (n.:35)	-0.26	-0.28	-0.35*	-0.36*	-0.35*	0.32	-0.30	-0.30	0.32	-0.30	-0.30	0.47**
Beech	0.48	0.11	0.11	-0.23	0.11	0.63*	0.53	0.27	0.53	0.35	0.35	0.71*
Oak	0.17	0.07	0.21	0.07	0.21	0.55	0.19	0.09	0.19	0.02	0.02	0.37
Chestnut	-0.46	-0.33	-0.40	-0.37	-0.40	0.48	0.20	-0.33	0.20	0.20	0.20	0.38
Grassland	-0.46	-0.24	-0.61	-0.42	-0.61	0.36	0.21	-0.49	0.21	-0.60	-0.60	0.26

** $\alpha < 0.01$, * $\alpha < 0.05$ (two-tailed test).

Correlation (r Pearson) performed between andic properties (Alo+0.5 Feo %) and NDVI metrics for each of the observed Land Cover classes (CORINE Land Cover classes) after site validation. The chemical analyses are integrated over soil depth (solum).