



1 **Revegetation in abandoned quarries with landfill stabilized waste and gravels:**
2 **water dynamics and plant growth—a case study**

3 Cheng-liang Zhang^{a**}, Jing-jing Feng^{bc**}, Li-ming Rong^a, Ting-ning Zhao^b

4 a. Beijing Key Lab of Industrial Land Contamination and Remediation,
5 Environmental Protection Research Institute of Light Industry, 100089, Beijing,
6 China;

7 b. School of Soil and Water Conservation, Beijing Forestry University, 100083,
8 Beijing, China;

9 c. School of Environmental Science and Engineering, Shanghai Jiaotong University,
10 Shanghai 200240, China

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13 Correspondence to: Tingning Zhao

14 Address: Beijing Forestry University, Qinghuadong Road 35, Beijing, China, 100083.

15 E-mail: zhtning@bjfu.edu.cn

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17 **These authors contributed to the work equally and should be regarded as co-first
18 authors.

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20 **ABSTRACT**

21 Large amounts of quarry wastes are produced during quarrying. Though quarry
22 wastes are commonly used in pavement construction and concrete production, in-situ
23 utilization during ecological restoration of abandoned quarries has its advantage of
24 simplicity. In this paper, rock fragments of 2~3 cm in size were mixed with landfill
25 stabilized waste (LSW) in different proportions (LSW: gravel, R_L), which was called
26 LGM. The water content, runoff and plant growth under natural precipitation were
27 monitored for two years using a runoff plot experiment. LGM with a low fraction of
28 LSW was compacted in different degrees to achieve an appropriate porosity; water
29 dynamic and plant growth of compacted LGM were studied in a field experiment. The
30 results showed that, (1) LGM can be used during restoration in abandoned quarries as
31 growing material for plants. (2) R_L had a significant effect on infiltration and water
32 holding capacity of LGM, and thus influenced retention of precipitation, water
33 condition and plant growth. LGM with R_L ranging from 8:1 to 3:7 was suitable for
34 plant growth, but the target species grew best when R_L was intermediate. (3)
35 Compaction significantly enhanced water content of LGM with a low R_L of 2:8, but
36 leaf water content of plants was lower or unchanged in the more compacted plots.
37 Moderate compaction was beneficial to the survival and growth of *Robinia*
38 *pseudoacacia*. *Platycladus orientalis* and *Medicago sativa* were not significantly
39 affected by compaction, and they grew better under high degree of compaction which
40 was disadvantageous for the uppermost layer of vegetation.

41 **Keywords:** abandoned quarry; rock fragment; compaction; municipal solid waste;



42 landfill



43 **1 Introduction**

44 During the process of civilization and urban construction, quarrying industry has fast
45 developed in order to satisfy the growing demand for stones (Duan et al., 2008).
46 Large amounts of quarry wastes are produced during quarrying. In 2006, the
47 byproduct of mining and quarrying accounted for 55% of industrial waste in Europe
48 (Castro-Gomes et al., 2012). Eighty percent of stones or soils extracted during
49 quarrying are wasted (André et al., 2014). Thirty percent and forty percent of marble
50 blocks are wasted as powder and rock fragments, respectively (Akbulut & Gürer,
51 2007). Quarry wastes occupy a lot of lands, have a low aesthetic value, and may lead
52 to soil erosion, landslide or debris flow.

53 Reusing is an environmental friendly way to deal with quarry wastes. Coarse and
54 fine aggregates are commonly used in quarry settlement, the production of concrete,
55 mortars and ceramic tiles, and highway construction (Amin et al., 2011; Safiuddin et
56 al., 2010, 2007). The coarse waste aggregates could also be used to make high
57 value-added products such as sculpture and architecture (Castro-Gomes et al., 2012),
58 while the ultra-fine marble wastes are used in various industrial processes such as
59 plastic, paper, pharmaceuticals industry and agriculture (Gazi et al., 2012). However,
60 the reuse of quarry waste on a national scale is primarily constrained by economic
61 aspects such as transport (Castro-Gomes et al., 2012) as well as the lack of industrial
62 symbiosis. Most quarrying and stone processing activities are performed by small
63 medium enterprises. This fragmental quarry industry hardly shares information,
64 services and by-product resources with other industrial practitioners in order to add



65 value, reduce costs, and improve environment, leading to inefficient waste
66 management (Gazi et al., 2012; Rubio et al., 2010). As stated by Tiruta-Barna et al.
67 (2007), the practitioners were lacking of unambiguous references concerning the
68 feasibility of specific reuse projects. Considering the current situation, in-situ
69 utilization of quarry waste during ecological restoration of abandoned quarries has its
70 advantage of simplicity.

71 Water and nutrient deficiencies are the main environmental factors constraining
72 natural recovery in abandoned quarries. These disadvantages site conditions have to
73 be changed to facilitate plant survival and growth (Luna et al., 2016). In the last
74 decade, many arable topsoil materials were bought and transported from other places
75 to replace or cover the barren spoil or waste in abandoned quarries or mines in China,
76 which was very expensive, accounting for about 50% of the total cost of restoration.
77 The environment of the places where surface soils are taken may also be adversely
78 affected. Soil-like materials containing high values of nutrients such as sewage sludge
79 and waste compost are effective topsoil substitutes because of their economic and
80 environmental advantages (Luna et al., 2016; Forján et al., 2016; Jordán et al., 2016).
81 Landfill stabilized waste (LSW) is the aged municipal solid waste which went through
82 a series of microbiological processes in the closed landfills. It is similar with
83 municipal waste compost in the source, production process and the properties of end
84 products. Previous research indicated its promising potential as growing substrate for
85 plants (Zhang et al., 2017; Zhou et al., 2015; Feng et al., 2017). However, compared
86 with sewage sludge and waste compost, LSW has not been fully exploited, and the



87 research on its properties and application is still limited.

88 When LSW is mined from the landfill, the original structure is inevitability
89 destroyed. With a high compressibility, LSW lacks of hydrostructural stability and is
90 very sensitive to compaction (Schäffer et al., 2008). On the contrary, quarry spoil and
91 waste may be highly compacted by machines or vehicles and thus have low hydraulic
92 conductivity. If LSW is directly placed on this impervious layer, slip surface may
93 generate between the two layers with dramatic difference in hydraulic conductivity or
94 within LSW layer because of the low shear strength of unconsolidated LSW (Okura et
95 al., 2003). As a result, instead of directly topsoiling with LSW, we mixed it with
96 gravels (2~3 cm in size) in different proportions, reusing rock fragments of quarry
97 wastes as part of the growing substrate for plants, which is called LGM hereinafter.
98 Not only the embedded gravels constitute the primary fabric and thus enhance the
99 stability of the mixture, the surface gravels can also reduce soil evaporation and
100 conserve surface water (Yuan et al., 2009), and prevent fine earth from wind and
101 splash erosion (Jomaa et al., 2012).

102 When the fraction of LSW is low and most macropores between rock fragments
103 are not filled with LSW, the water holding capacity and thus the water condition may
104 be disadvantageous for plant growth. In order to achieve an appropriate porosity,
105 LGM with a low LSW: gravel ratio (2:8) was compacted in different degrees. The
106 effects of compaction on water condition and plant growth were also studied.

107 **2 Materials and methods**



108 **2.1. Study area**

109 The research was conducted in the Ecological Restoration Research Base of
110 Environmental Protection Research Institute of Light Industry, located in Changping
111 County, Beijing (40°9'56.73"N, 116°9'1.04"E, 57 m a.s.l.). Beijing has a continental
112 monsoon climate with a rainy season from June to September. The mean annual
113 precipitation is 600 mm, the mean annual temperature is 8~12°C, and the mean annual
114 evaporation is 1800~2000 mm.

115 **2.2. Materials**

116 LSW was taken from five landfills in different districts of Beijing, transported to the
117 study site, and mixed in 2012. Texture, chemical properties such as nutrient and heavy
118 metal contents, concentrations of semi-volatile contaminants and MPN of coliforms
119 were tested. The texture of LSW was in accord with sandy clay loam; generally, the
120 nutrient contents of LSW were higher than the local agricultural soil; the
121 concentrations of heavy metals and organic contaminants were lower than the national
122 limits or similar to the concentrations in natural soil. Basic physical and chemical
123 properties are shown in Tab.1.

124 Gravels were bought from local market. The mean values (with standard
125 deviation) of maximum, medium and minor axis length were 31 ± 24 , 21 ± 4 and 15 ± 4
126 mm respectively. LSW and gravels were mixed with a mixing machine in different
127 fractions. The LSW: gravel ratios by volume (R_L) were 8:1, 7:3, 5:5, 3:7 and 1:8,
128 which were marked as L1, L2, L3, L4 and L5. LGMs were oven-dried, weighted, and
129 filled into 660-cm³ cubic flowerpots. They were saturated and weighted to calculate



130 porosity (saturated water content), and then allowed to drain freely under gravity for
131 one day and weighted to calculate field capacity. The upper openings of the
132 flowerpots were sealed with plastic film to prevent evaporation, and gauze was used
133 to avoid leakage of the finer grains through the drain holes.

134 **2.3 Runoff plots of LGM**

135 **2.3.1 Setting of runoff plots**

136 Five 2.8-m-wide, 3.6-m-long runoff plots with a slope degree of 38 ° were set in 2012.
137 70-cm-thick LGM (L1~L5) was spread on the impervious liners and was settled for 1
138 year. Ruderal species such as *Setaria viridis*, *Digitaria sanguinalis*, *Amaranthus*
139 *retroflexus*, *Bidens parviflora*, *Pharbitis nil*, *Echinochloa crusgalli*, *Chenopodium*
140 *album* and *Metaplexis japonica* colonized during the 1-year settlement. During the
141 rainy season the vegetation coverages from L1 to L4 were 100% while that of L5 was
142 about 50%. Litters in two 1-m² quadrats for each runoff plot were collected in
143 November 2014 to evaluate the abundance of ruderals.

144 **2.3.2 Water content**

145 The volumetric water content at 10~50 cm depths was measured using a capacitance
146 probe (Diviner 2000, Sentek Pty Ltd., Australia) from August 2013 to August 2015,
147 three times a month; these data were averaged to calculate annual and monthly mean
148 water content. The volumetric water content was also measured every sunny day from
149 May to August 2014; these data were used to study the relationship between soil
150 moisture and precipitation during the rainy season.



151 2.3.3. Retention of precipitation

152 Multiple regression model as follows is fitted:

$$153 Y = \alpha + \beta_1 X_1 + \beta_2 X_2 \quad (1)$$

154 where X_1 , X_2 and Y represent antecedent water content (%), precipitation (mm) and
 155 water content one day after the rainfall event (%), respectively. The partial regression
 156 coefficient for precipitation (β_2) is used as a measure of water conservation, i.e., the
 157 average increment of water content each additional precipitation is associated with,
 158 for any given antecedent water content (Cohen et al., 2003). Retention of precipitation,
 159 which is the percentage of precipitation able to infiltrate and stored in the LGM
 160 profile after one night's drainage is calculated as follows:

$$161 RP = \beta_2 \times D \quad (2)$$

162 where RP is the retention of precipitation, and D is the depth of LGM.

163 2.3.4 Runoff

164 Surface and subsurface runoff were collected separately one day after each rainfall
 165 event from June to September (the rainy season) in 2013 and 2014. During the
 166 experimental period, there were altogether 32 rainfalls, including 12 light rains (<10
 167 mm d^{-1}), 15 moderate rains ($10\sim25 \text{ mm d}^{-1}$), and 5 heavy rains ($>25 \text{ mm d}^{-1}$).

168 2.3.4 Plant growth

169 After clearing the colonized ruderals, seed mixture of *Robinia pseudoacacia*
 170 (leguminous tree, the target species), *Festuca elata* (perennial grass),
 171 *Orychophragmus violaceus* (annual or biennial herb), *Viola philippica* (perennial herb)
 172 was sowed with a density of 15 g m^{-2} (6:2:1:1 by mass) in June 2015. Former study



173 indicated its efficiency in fast revegetation and providing a stable soil cover (Feng et
174 al., 2015). Vegetation coverage was measured from August to November 2015 using
175 eight 1-m² fixed sample plots (Xie, 2010). Two tallest *R. pseudoacacia* seedlings in
176 each sample plots were cut in November 2015, oven dried for 12 h, and the biomass
177 was measured.

178 **2.4 Compacted plots of LGM**

179 **2.4.1 Setting of compacted plots**

180 Five 5-m-long, 3-m-wide, 1-m-deep plots were dug and filled with the LGM with a
181 low LSW: gravel ratio of 2:8 ($R_L=2:8$). LGM was compacted using a vibratory roller
182 in different degrees.

183 **2.4.2 Determination of theoretical porosity**

184 Because the porosity of compacted LGM was difficult to attain, it was calculated
185 theoretically, assuming that the mass and particle density of the loose LSW were
186 constant:

$$187 \quad \rho = M / V_0 (1 - P_0) = M / V_c (1 - P_c) \quad (3)$$

188 where ρ is particle density; P_0 or P_c is the porosity of LGM before or after
189 compaction; M is the mass of LGM particle; V_0 or V_c is the volume of LGM before or
190 after compaction. Hence,

$$191 \quad P_c = 1 - (1 - P_0) \times V_0 / V_c \quad (4)$$

192 P_0 was measured as described in Section 2.2; V_0 and V_c were measured before
193 and after the compaction.



194 **2.4.3 Water content and retention of precipitation**

195 The volumetric water content at 10~100 cm depths was measured from three PVC
 196 tubes in each compacted plot using the capacitance probe. The measurement of water
 197 content and the calculation of RP in the compacted plots were the same with the
 198 measurement and calculation in the runoff plots.

199 **2.4.4. Plant growth**

200 56 *R. pseudoacacia* (leguminous tree) and 30 *Medicago sativa* (leguminous herb)
 201 seeds were sowed and 60 *Platycladus orientalis* (evergreen conifer) seedlings were
 202 transplanted in each compacted plot after compaction in 2012. No water or nutrients
 203 were applied. Germination rate was measured in October 2012 and survival rate was
 204 measured in October 2014. From 2013 to 2014, three times a month during rainy
 205 season, leaves were collected at 12 a.m., weighted, oven-dried and weighted again to
 206 calculate leaf water content; height and stem base diameter of each woody seedling
 207 were measured in October; 15 individuals of *M. sativa* were cut after flowering,
 208 oven-dried and the aboveground biomass was measured.

209 **2.5. Data analysis**

210 Data were log-transformed and normalized when necessary. One-way analysis of
 211 variance and least significant difference test were used to compare height and biomass
 212 of *R. pseudoacacia* seedlings growing in LGMs with different R_{LS} . This test was also
 213 used to compare leaf water content, height or diameter growth (for trees) and biomass
 214 (for herbs) of seedlings growing in different compacted LGM plots. Friedman test
 215 was used to compare water content, runoff and vegetation coverage of different runoff



plots. Linear regression model was used to describe the relationship between antecedent water content, precipitation and water content. SPSS software was used for data analysis.

3 Results

3.1 LGM with different fractions of LSW

3.1.1 Water content and use efficiency of precipitation

As shown in Tab.2, the field capacity of LGM decreased significantly with decreasing R_L ($P < 0.01$), reflecting a positive relationship between water holding capacity and R_L since most capillary pores were provided by LSW. The saturated water content was lowest when R_L was intermediate and LSW exactly filled the voids between the gravels, but the difference was not significant.

The annual mean, maximum and minimum monthly mean water contents of all LGM plots were much lower than field capacity, indicating prolonged soil water deficit. However, plants may response quickly to pulsed rainfall events and make the best of precipitation (Huxman et al., 2004). R_L had a positive effect on retention of precipitation. With a decreasing R_L , the percentage of precipitation which was able to infiltrate and was stored in the LGM profile dropped from 70.5% to 24.5%.

3.1.2 Runoff generation

Under light rains, surface or subsurface runoff did not change significantly with R_L , but the volume of total runoff in L5 was significantly higher than those in L2 and L3 ($P < 0.05$). During moderate rainfalls, the surface runoff in L2 was significantly lower than those in L3, L4 and L5 ($P < 0.05$), but subsurface or total runoff did not change



significantly with R_L . During heavy rainfalls, surface, subsurface or total runoff did not change significantly with R_L . However, as shown in Fig.1, generally, during moderate and heavy rainfalls, subsurface and total runoff had a tendency to increase with decreasing R_L . The failure in passing the significance test may result from the high variation of runoff under natural precipitation, indicated by its high standard deviation, which may be caused by the different intensity and/or duration of each rainfall event and the antecedent water content (Liu et al., 2012).

3.1.3 Plant growth

From August to December, vegetation coverage from L1 to L5 were 28.9%~86.0%, 20.2%~96.0%, 37.2%~93.4%, 24.8%~82.3% and 6.2%~43.0%, respectively. Generally, the speed of vegetation formation, the mean coverage and the duration of land cover were similar from L1 to L4, but L5 showed distinctly lower values. Height and biomass of *R. pseudoacacia* were significantly higher in L3 ($P<0.01$); the differences were not significant between L1, L2 and L4; L5 still had the lowest performance (Tab.3).

3.1.4 Litter

The dry weights of litter from L1 to L5 were 0.283, 0.257, 0.197, 0.217 and 0.086 kg m⁻², respectively, which showed a decreased tendency with decreasing R_L .

3.2 Compacted LGM with low fraction of LSW

3.2.1 Water condition of compacted LGM

As shown in Tab.4, the annual mean water content of LGM ($R_L=2:8$) increased significantly with increasing degree of compaction ($P<0.01$). The theoretical porosity



260 was 13.3%~26.1% higher than the annual mean water content in CL1~CL4, but 2.1%
261 lower in CL5, indicating its poor aeration, which may hinder microbial activity,
262 nutrient mineralization, and the uptake of water and nutrient by plants.

263 Compaction had a positive effect on retention of precipitation. With an
264 increasing compaction degree, the percentage of precipitation which was able to
265 infiltrate and was stored in the LGM profile increased from 34% to 97%.

266 3.2.2 Plant growth in the compacted LGM

267 3.2.2.1 Leaf water content

268 Leaf water content of *R. pseudoacacia* and *M. sativa* decreased significantly with
269 increasing degree of compaction ($P < 0.05$). But the difference of *P. orientalis* was not
270 significant between different compacted plots (Tab.5).

271 3.2.2.2 Survival rate

272 The germination rates of *R. pseudoacacia* from CL1 to CL5 were 95%, 87%, 92%,
273 64% and 34% respectively in 2012; the survival rates were 59%, 56%, 56%, 28% and
274 31% respectively in 2014.

275 The survival rates of *P. orientalis* from CL1 to CL5 were 100%, 97%, 97%, 95%,
276 and 95% respectively in 2012 and were 85%, 83%, 83%, 88% and 88% respectively
277 in 2014.

278 3.2.2.3 Growth rate

279 Compaction had a significant suppressive effect on height or diameter growth of *R.*
280 *pseudoacacia* ($P < 0.01$), but the height and stem base diameter of *P. orientalis*
281 increased with an increasing degree of compaction (Fig. 2).



282 The effect of compaction on biomass of *M. sativa* was positive and significant
283 ($P < 0.01$). The difference between CL1 and CL2 was not significant, but the biomass
284 was significantly higher in CL3~CL5 compared to CL1 in both 2014 and 2015 (Fig.3).

285 4 Discussion

286 4.1 Using LGM as growing substrate for plants

287 LSW is a soil-like material containing high values of organic matter and other
288 nutrients such as nitrogen and potassium, and thus has a promising prospect as topsoil
289 substitute during environmental restoration. However, unconsolidated LSW is
290 unstable and prone to wind or water erosion especially when its structure is destroyed
291 during landfill mining. When LSW is mixed with rock fragments, friction force and
292 interlock capacity would increase while surface runoff and soil erosion would
293 decrease (Descroix et al., 2001). Compared to LSW, less surface runoff was generated
294 in LGM regardless of R_L or precipitation intensity. The effect of R_L on surface runoff
295 was not significant, probably because the infiltration rate was always higher than the
296 intensity of precipitation, and thus most rainfall infiltrated into LGM profile, held by
297 capillary force or discharging as subsurface runoff. Though the water content of LGM
298 was only 15.4%~50.9% of LSW under natural precipitation (Zhang et al., 2017),
299 plants grew well in L1~L4. R_L significantly influenced water holding capacity of
300 LGM, which was reflected by the positive relationship between R_L and field capacity
301 and the negative relationship between R_L and subsurface runoff. More water was
302 retained in the LGM profile with a higher R_L and available for plant use, thus
303 facilitating plant growth. However, the target species, *R. pseudoacacia* grew best in



304 L3, which may result from the less severe inter-specific competition compared to L1
305 and L2. A higher contents of nutrients provided by LSW may be more beneficial to
306 ruderals and other herbaceous species (Le Stradic et al., 2014), which was consistent
307 with the higher amount of ruderal litters in L1 and L2. L5 showed the poorest water
308 condition and plant growth because water cannot be held within macropores between
309 gravels, and thus LGM with this R_L is not recommended unless other ameliorative
310 measures are supplemented.

311 **4.2 Using compacted LGM as growing substrate for plants**

312 Compaction had a significant effect on improving water holding capacity. With an
313 increasing compaction degree, the annual mean water content increased from 5% to
314 15.4%, and retention of precipitation increased from 34% to 97%. However, leaf
315 water contents of *R. pseudoacacia* was significantly lower in the compacted plots,
316 indicating water deficit. *R. pseudoacacia* is a fast-growing species with deep root
317 system. The high transpiration demand was not satisfied probably for two reasons:
318 firstly, soil water in the compacted plots may be harder to extract because of lower
319 water potential and hydraulic conductivity; secondly, water uptake may decrease
320 because of the constrained development of roots (Sharrow, 2007; Nadian et al., 1996).

321 An unneglectable deficiency of the experiment was that plant roots could grow
322 through the 1-m-thick LGM layer and absorb water from the underlying natural soil
323 because impervious liner was not set between LGM and the underlying soil. Plants
324 growing in the uncompacted LGM in real environment i.e. abandoned quarries,
325 should not have this good performance because the only water source was LGM



326 which contained only 5% water under natural precipitation. *R. pseudoacacia* growing
327 in CL3 should have the best performance in the real environment although the
328 germination, survival and growth were not significantly different from CL1 in our
329 experiment, which was consistent with Jeldes et al. (2013).

330 Leaf water content of *P. orientalis* was not affected by compaction, and seedlings
331 grew better in the more compacted plots, indicating that water or compaction degree
332 was not the key factor constraining growth rate. Many researches have shown that,
333 conifers have a more conserving water use strategy compared to board-leaved tree
334 species (Catovsky et al., 2002; Wullschleger et al., 1998) and evergreens are more
335 adapted to compacted soils than deciduous (Alameda & Villar, 2009). However, light
336 intensity may have a significant effect on *P. orientalis*. As the canopy density of *R.*
337 *pseudoacacia* decreased with increasing degree of compaction, more light was
338 available to *P. orientalis* which grew underneath, improving its performance.

339 Leaf water content of *M. sativa* decreased while biomass increased with
340 increasing degree of compaction. *M. sativa* was well adapted to compacted plots,
341 which was consistent with Cresswell & Kirkegaard (1995). The higher biomass may
342 result from less intensive interspecific competition for resources such as light, or
343 because plant growth was accelerated under mild or temporary water stress (Shao et al.
344 2010).

345 5 Conclusion

346 LGM, the mixture of landfill-stabilized waste and coarse quarry waste can be used
347 during restoration in abandoned quarries as growing material for plants. The LSW:



348 gravel ratio had a significant effect on the physio-chemical properties such as nutrient
349 level, water condition, physical stability, and thus the performance of plants growing
350 in LGM. LGM with R_L ranging from 8:1 to 3:7 was suitable for plant growth, but the
351 target species grew best when R_L was intermediate. When R_L was lower than 3:7,
352 compaction enhanced the retention of precipitation, but leaf water content of plants
353 was lower or unchanged in the more compacted plots. Moderate compaction was
354 beneficial to the survival and growth of *R. pseudoacacia*, *P. orientalis* and *M. sativa*
355 were not significantly affected by compaction; they grew better in highly compacted
356 area where the uppermost layer of vegetation was suppressed and thus more light was
357 available. Compared to fast-growing broad-leaved trees, conifers and herbaceous
358 species may be more adapted to compacted LGM, and interspecific interaction
359 showed a significant effect on plant performance. Nutrient or pollutant leaching with
360 the deep percolation water or surface run-off are not covered in this paper, but they
361 are very important factors considering the long-time impacts of LGM or LSW
362 application. Some work has been done on the chemical and hydrological properties of
363 LSW (Zhang et al., 2017; Feng et al., 2017) but more study should be taken on the
364 environmental effects of municipal solid waste application.

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475



476 Tab.1 Basic physical and chemical properties of LSW (\pm standard deviation)

Objects	Mean values	Objects	Mean values
Sand (0.05~2 mm, %)	64	As (mg kg ⁻¹)	10.1 \pm 5.5
Slit (0.002~0.05, %)	16	Cr (mg kg ⁻¹)	91.2 \pm 46.4
Clay (<0.002 mm, %)	20	Cu (mg kg ⁻¹)	77.6 \pm 51.6
Total K (g kg ⁻¹)	25.5 \pm 7.7	Ni (mg kg ⁻¹)	31.9 \pm 12.4
Available K (mg kg ⁻¹)	503 \pm 124	Pb (mg kg ⁻¹)	54.2 \pm 30.8
Total N (g kg ⁻¹)	1.95 \pm 0.51	Zn (mg kg ⁻¹)	215 \pm 36.1
Nitrate-N (mg kg ⁻¹)	105.9 \pm 105.1	Cd (mg kg ⁻¹)	0.29 \pm 0.05
Ammonium-N (mg kg ⁻¹)	9.21 \pm 4.78	Hg (mg kg ⁻¹)	0.78 \pm 0.45
Total P (g kg ⁻¹)	1.12 \pm 0.33	pH value	8.2
Available P (mg kg ⁻¹)	75.08 \pm 145.87	Organic matter (g kg ⁻¹)	40.12 \pm 22.27

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478



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Tab.2 Hydrophysical properties of LGM

LSW fraction	Moisture content (%)					RP (%)
	Saturated	Field	Annual	Maximum	Minimum	
	moisture content	capacity	mean	monthly mean	monthly mean	
L1	48.9±2.0	42.9±2.0	8.7±1.6	10.9	6.8	70.5
L2	39.2±10.8	33.2±10.8	4.8±0.8	5.8	3.6	34.0
L3	31.6±0.1	25.7±0.1	5.2±1.3	7.6	3.8	49.0
L4	34.3±4.1	13.6±3.4	3.2±0.7	4.5	2.3	26.5
L5	41.9±1.2	8.7±1.2	2.6±0.8	3.9	1.9	24.5

480

Note: The multiple regression fitted to attain RP was significant at the 0.01 level. Antecedent

481

water content and precipitation accounted for 73.4%~87.6% variance of water content 1 d after the

482

rainfall event.

483



484

Tab.3 Height and single plant biomass *R. pseudoacacia*

Runoff plot	Height (cm)	Above ground biomass of single plant (g)
L1	22±6 ^b	0.759±0.159 ^b
L2	25±8 ^b	0.958±0.317 ^b
L3	35±6 ^a	2.035±0.480 ^a
L4	26±7 ^b	0.917±0.095 ^b
L5	11 ± 4 ^c	0.171±0.031 ^b

485

Note: the same letter indicates that the difference is not significant at the 0.05 level.

486


487 Tab.4 Hydro-physical properties of LGM ($R_L=2:8$) with different compaction degrees

Compacted plot	Moisture content (%)				RP (%)
	Theoretical	Annual	Maximum monthly	Minimum monthly	
	porosity	mean	mean	mean	
CL1	31.1	5.0 ± 0.8	6.1	4.1	34
CL2	26.8	6.4 ± 1.2	8.0	5.0	45
CL3	22.1	7.8 ± 1.2	9.4	6.5	44
CL4	21.1	7.7 ± 1.2	8.9	6.4	53
CL5	13.3	15.4 ± 2.1	17.8	13.4	97

488 Note: the multiple regression fitted to attain RP was significant ($P < 0.01$). Antecedent water
489 content and precipitation accounted for 65.7%~86.1% variance of water content 1 d after the
490 rainfall event.

491



492 Tab.5 Average leaf water contents of plants growing in the compacted LGM (g g^{-1})

Plant species	CL1	CL2	CL3	CL4	CL5
<i>R. pseudoacacia</i>	2.25 ± 0.09^A	2.19 ± 0.06^A	2.20 ± 0.13^{AB}	1.99 ± 0.06^B	1.88 ± 0.09^B
<i>P. orientalis</i>	2.04 ± 0.17	2.16 ± 0.14	2.10 ± 0.16	2.07 ± 0.15	2.12 ± 0.17
<i>M. sativa</i>	3.79 ± 0.15^A	3.50 ± 0.18^A	3.29 ± 0.19^A	3.13 ± 0.22^{AB}	2.45 ± 0.18^B

493 Note: the same letter indicates that the difference between different compacted plots is not
 494 significant at the 0.05 level.

495

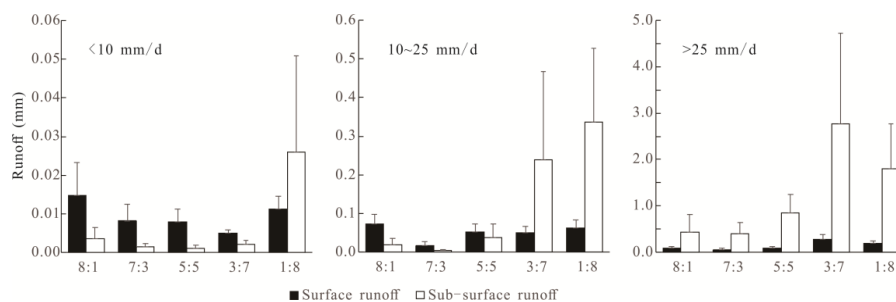
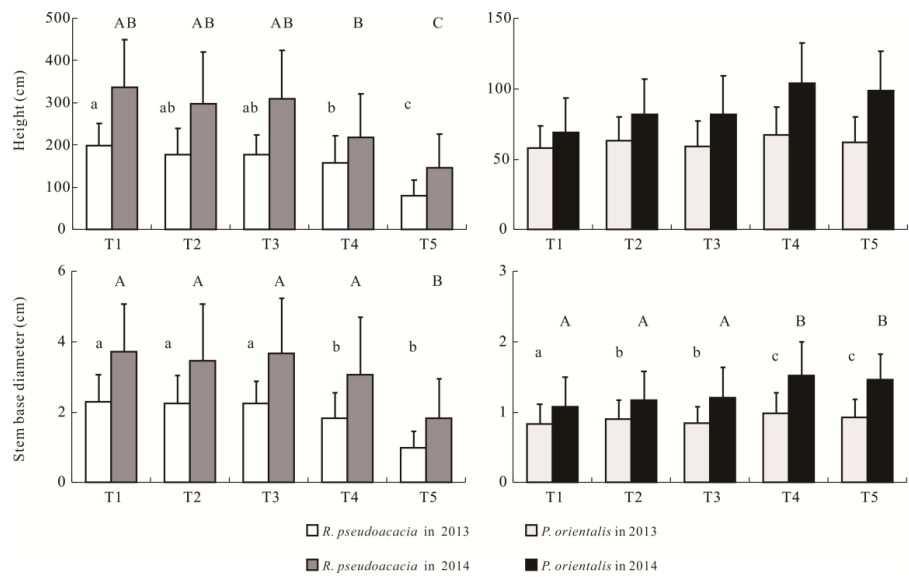


Fig.1 Average runoff generated in LGM slopes under natural precipitation



500

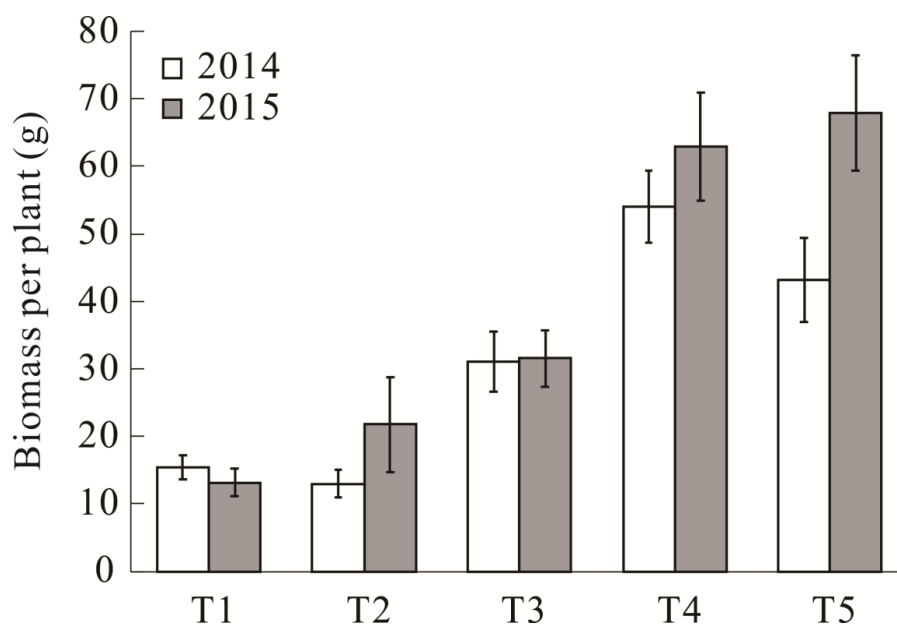
501 Fig.2 Wood growth

502 Note: the same small or capital letter indicates that in 2013 or 2014 the difference between

503 different compacted plots is not significant at the 0.05 level.

504

505



506

507 Fig.3 Herb growth

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